Report on the Environmental Benefits

Environmental Benefits of Recycling – 2016 edition

Bureau of International Recycling (BIR)

BIR-Nominated Commodities: Aluminium, Copper, Ferrous and Paper



Report on the Environmental Benefits of Recycling – 2016 edition
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Executive Summary

Use of a more detailed and refined methodology reveals that carbon dioxide emission savings achieved through recycling as compared to primary production are greater than concluded by the 2008 BIR report which arrived at a figure of around 500 million tonnes per annum for seven metals (ferrous, aluminium, copper, nickel, tin, zinc and lead) plus paper — equivalent at the time to the annual carbon dioxide emissions of the global aviation industry.

The latest in-depth study conducted on behalf of BIR concludes that, for just three metals (ferrous, aluminium and copper), the annual ${\rm CO}_2$ savings made through secondary production rather than primary production is 572 million tonnes (see table). The report also highlights the energy and emissions benefits of recycling paper.

The report supplements literature-based and benchmark data from the 2008 study with "real data" derived from industry, including use of a novel "Front-end" tool for "normalisation" of industry-acquired data in terms of energy requirements and associated ${\rm CO}_2$ emissions.

The methodologies described in the report can be used to obtain potential CO_2 savings for any company's recycling operations for any material on a case-by-case basis.

Material	Energy Savings (achieved by industry against Primary Benchmark) (TJ/100,000t)	Annual Worldwide Secondary Production* (Mt)	Estimated Savings in Annual CO ₂ Emissions (Mt)
Aluminium	4434	18	63.3
Copper	1033	6	4.8
Ferrous	206	580	503.9
	in Annual CO ₂ Emissions econdary Metals Studied [Current Studied]	dy]	572.0

^{*} Annual worldwide secondary production (Mt) as quoted in 2014 for Aluminium and in 2013 for Copper and Ferrous.

The benefits of the 2008 research, a desk-based study of the available literature of the primary and secondary production of seven metals and paper which made use of the best available and most justifiable energy data for primary and secondary production in calculations to highlight the carbon footprint savings associated with secondary production, have been widely acknowledged in the recycling industry worldwide. It is in this context that the current study is carried out: "... to update the findings of the 2008 study, applying the methodology developed to more recent data on four of the original commodities, namely, aluminium, copper, ferrous metals and paper".

In the current study use was made of industryacquired data drawn from real secondary recovery operations to improve the input data for secondary production and, as in the 2008 study, to avoid complications associated with the early stages of the whole lifecycles of the materials benchmark, energy requirements and carbon footprints were taken from ore or raw material delivered, respectively, at the production plant for primary material, and at the production plant for secondary material. The data provided by industry permit calculations of energy requirements and carbon dioxide emissions for both single-stream processes for each of the BIRnominated commodities, and separately, for mixedmetal streams containing all three target metals. For the single-stream process, the data provided allow for calculations to be made for the recovery of a fully-refined product for each commodity whilst for the mixed-stream process, the data allow for calculations to be made for the recovery of a

"saleable" product which would require further refinement to a fully-refined product.

To realise the full potential and value of applying the methodology used in this study, industry engagement is important so that the energy requirements and carbon footprints are determined on industry-acquired data (i.e. "real" data). To this end, a novel "Front-end" tool has been developed for "normalisation" of industry-acquired data (in terms of energy requirements and associated CO₂ emissions), as input to the methodology. To optimise the value of the output from the analysis, the nature and type of information required from industry on recycling operations form the basis of a questionnaire developed to assist BIR in the acquisition of information from its members and other stakeholders. In situations where industry is handling mixed streams from which value can be derived from more than one commodity, account is taken of the energy apportioned to each metal and non-metal fraction. Use of the "Front-end" tool allows fractionation and attribution of energy data in the process and for the energy and CO₂ emissions for each recovered commodity to be determined. In situations where there is less than 100% recovery of useful material, the tool can be extended to attribute that energy to further recovery of the product(s).

The CO₂ emissions for aluminium, copper, ferrous metals and paper are calculated using an electricity energy conversion factor, for the UK, of 0.50935kgCO₂e/kWh. Energy conversion factors vary considerably between countries and regions and the CO₂ emissions can be re-calculated for any specific country or region by using the appropriate conversion factor. The effects of other variables,

for example, plant and operation efficiencies as well as energy source/fuel mix, can be determined using the sensitivity analyses tables provided.

The CO_2 emissions savings achieved by secondary production over primary production for the industry-derived data are presented in the following sets of comparisons:

- 1 For the metals and paper, with a calculation from benchmark primary energy data, from the 2008 study, using the 0.50935kgCO₂e/kWh electricity energy conversion factor, assuming that primary and secondary production can be carried out in the same geographical region where this conversion factor is appropriate. The comparative calculation can be extended to processes carried out in any region or country by use of the relevant conversion factor; and
- 2 For the metals only, with the benchmark primary calculations from the 2008 study representing the most efficient production processes available with the lowest energy consumption in situations where the best possible energy mixes are used anywhere in the world, using the most recent worldwide secondary production tonnages for these metals, the total estimated savings in annual CO₂ emissions arising from the secondary production of aluminium, copper and ferrous metals, in comparison with primary production, are 572Mt.

Furthermore, the methodology described in this report can be extended and used to obtain energy and carbon emissions data on any recycling operations of any process operator and for any material.

Context of the Brief

Background

The environmental benefits of recycling have been expressed in many ways, including savings in energy and in use of virgin materials. Very little attempt had been made, however, to express these benefits in terms of carbon footprint and particularly as savings in CO_2 equivalent emissions, which would have implications in terms of both the environment and carbon emissions, until BIR commissioned a research study on behalf of its members in 2008:

"...to prepare a report on the environmental benefits of recycling, identifying the savings that can be made by using recyclables as opposed to primaries, and thereby the carbon credentials of the recycling industries".

The research, directed by Professor Sue Grimes, was presented as a lead industry document, entitled *Environmental Benefits of Recycling*, referred to in this work as the '2008 BIR report'.

This desk-based research involved a detailed review of available scientific and technical literature on seven metals – aluminium, copper, ferrous metals, lead, nickel, tin and zinc – and also paper. The study introduced the concept and application of a benchmark methodology to determine the best available and most justifiable carbon emission data for primary production processes, and used best estimates from the literature of benchmark data for energy and carbon footprint calculations for both primary and secondary production. For primary production, the benchmark data represented the most efficient production processes available with the lowest energy consumption per tonne of metal produced in situations where the best possible energy mixes were used. The conversion factors used to express the primary production energy data as benchmark carbon emission data were also based on those for the best possible energy mixes. Benchmark data were, thus, defined as those data that represented material production situations that were achievable and gave values that were most acceptable and justifiable as the best achievable, but would not necessarily be achieved by all

primary production processes. The calculations of benchmark values for secondary production in the 2008 study were similarly derived from literature-based data.

The benchmark data were used to highlight the advantages (environmental impacts) of secondary production over primary production and were reported per 100,000 tonnes of material produced to provide a means of direct comparison between primary and secondary production and expressed as CO₂ savings per 100,000 tonnes of production. To avoid complications associated with the early stages of the whole lifecycles of these materials, benchmark energy requirements and carbon footprints were taken from ore or raw material delivered at the production plant for primary material, and delivered at the secondary plant for secondary material. Sensitivity analyses were then developed and used to show how the comparisons can be handled to deal with variations in different production processes, for example, variations in efficiency, and fuel and energy balances.

The concept of benchmarking was a novel approach to calculating environmental parameters such as carbon footprints and carbon dioxide savings and, combined with the provision of sensitivity analyses, provides a means of obtaining the best available calculated data for individual situations.

The benefits of the 2008 research have been widely acknowledged in the recycling industry worldwide, and, for the recycling industries, the value of expressing environmental benefits in terms of ${\rm CO_2}$ emissions savings is becoming increasingly necessary. It is in the context of the success of this work that Alexandre Delacoux, Director General of BIR, approached Professor Sue Grimes to consider carrying out further research on behalf of BIR:

"... to update the findings of the 2008 study, applying the methodology developed to more recent data on four of the original commodities, namely, aluminium, copper, ferrous metals and paper".

Scope and Assumptions

The scope outlined for the current report is based on the above brief.

In carrying out the project, the following assumptions have been made:

- 1 The scope for determination of carbon and energy savings will focus on four of the original commodities, as proposed by BIR, and referred to in this report as the BIR-nominated commodities: aluminium, copper, ferrous metals and paper;
- 2 The update of the 2008 study, applying the methodology developed to more recent data, will make use of primary and secondary literature as information sources, supplemented by published data from commodity associations and data that can be provided by BIR or other industry sources. Data on the material and energy balances provided by industry will be used, where available, to determine the carbon status of the recovery process of each material under study and will be compared with data obtained on primary material production;
- 3 The key deliverables will include:
 - a) A report setting out the results of CO₂ emissions savings and energy savings updated (from most recent data) for the BIRnominated commodities;
 - b) A list of questions/key indicators for BIR to present to its members to encourage their support for and participation in the industryspecific aspects of the study to permit the assessment of CO₂ emissions savings and energy savings applying the methodology using verifiable industry-acquired data for specified materials, specified plants in specified countries; and
 - c) Support to BIR in disseminating the results of the study to its members and other stakeholders, in the form of headline nontechnical statements, based on the scientific findings of the study.

To avoid complications associated with the energy usage and carbon footprints of the early stages of primary and secondary production, the 2008 study compared data across production plants based on: (i) ore concentrate delivered to the primary process facility converted to final product; and (ii) scrap and secondary material delivered to the secondary recovery facility converted to final product.

This assumption provides the most direct comparison between the environmental impacts of primary and secondary production because it focuses on the part of the life cycle that deals only with the actual production process and avoids problems associated with assessing the effects of mining and beneficiation of primary ores and with the collection and transport/delivery of scrap. It is likely however that the energies and carbon footprints associated with mining and concentration of ores would be much larger than collection and delivery of scrap and this would only serve to increase the benefits of secondary recovery.

It should be noted that the benchmark carbon footprints from the 2008 study essentially assume the use of the most appropriate and best available technologies. The assumptions made in the 2008 study will be upheld in the current study with the unit of comparison of 100,000 tonnes of produced material applied.



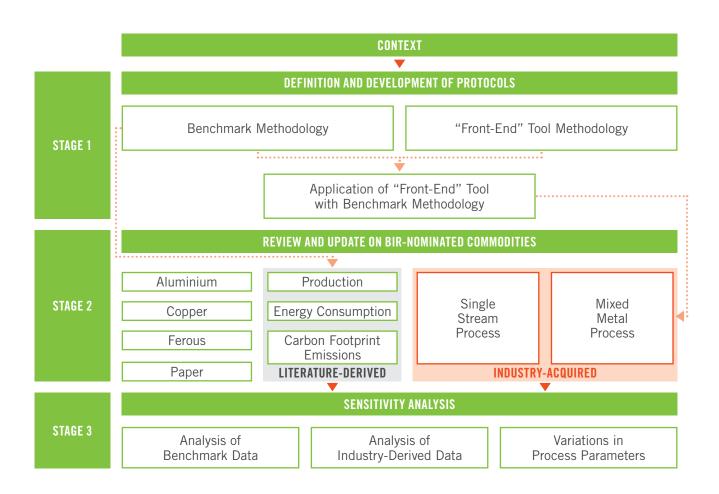
Approach

To achieve the key deliverables in line with the scope of the brief, the work is carried out in three stages, as illustrated in the schematic below, such that the following objectives are met:

- Agree the scope and deliverables of the brief with BIR;
- For the purposes of benchmarking, carry out a review of relevant primary literature on the materials under study;
- Obtain where possible process data from industry for use in carbon and energy consumption calculations;
- Compare the data from known secondary recovery operations with benchmark values for primary production, which represent the best available and most justifiable carbon data;
- Assess the quality of the data in terms of sensitivity analysis;

- Analyse and interpret the findings;
- Prepare headline non-technical statements, based on the study;
- Produce a draft final report for review and acceptance by BIR; and
- Provide a list of questions/key indicators for use by BIR in encouraging member participation in the industry-specific aspects of the study.

The focus of the current study is designed to update the 2008 work, with emphasis on the three metals aluminium, copper and ferrous and on paper, extended to use industry-acquired data to calculate the energy requirements and carbon footprint data of each of the commodities; allow comparison of the energy-specific data with the 2008 benchmark values for primary and secondary production; and apply sensitivity analyses to the industry-acquired data to take account of other situations such as different efficiencies in a plant.



Definition and Development of Protocols

The most common greenhouse gas emitted is carbon dioxide and a carbon footprint is a quantitative measure of the carbon dioxide released as a result of an activity expressed as a factor of the greenhouse gas effect of carbon dioxide itself. Many environmental impacts, including the production of any electricity used in the materials recovery industry, can be converted into carbon dioxide-equivalent (CO₂-e) emissions.

Benchmark Methodology

As a basis for reference, the benchmark methodology developed in the 2008 study, and applied here, involves:

- 1 A detailed survey of the literature to extract the data available on energy consumption in primary and secondary material recovery and the carbon emissions associated with these processes;
- 2 Use of energy data and associated carbon emissions, extracted to highlight differences between primary and secondary production of the BIR-nominated commodities. The assumptions made in all information provided are identified and the units used in the calculations are expressed as MJ/kg (or GJ/kg) of product for energy and tonnes of CO₂ per tonne of product for carbon emissions;
- 3 For each material for both primary and secondary production, best estimates of benchmark energy consumptions and carbon footprints are used in the comparisons as examples of what can be achieved;
- 4 A comparative analysis of energy consumption and carbon footprint data from the primary and secondary production of the BIR-nominated commodities, expressed per 100,000 tonnes of product. For all materials, the life cycle boundaries are set to compare the production of (a) primary material from raw material extraction to production of final concentrate/ product, and (b) recovered secondary materials ready for delivery to the recycling plant to final product. This enables a direct comparison to be made between the energy consumption required to produce primary and secondary refinery feedstocks for the same quantity of contained recyclate; and

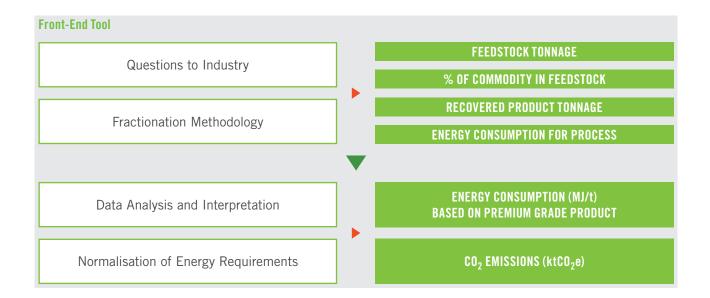
5 Sensitivity analyses are carried out on the data obtained using the benchmark values derived from the study to show how these data can be modified/handled to deal with variations in input such as energy sources used, energy/fuel mix for different countries, and energy efficiency of specific recovery plants.

Building on the acknowledged benefits of the 2008 research, it is clear that to realise the full potential and value of applying the benchmark methodology relies on industry engagement such that the energy requirements and carbon footprints are determined based on industry-acquired data (i.e. "real" data). Of further benefit is the ability to apply sensitivity analyses to these "real" data to take account of differences in, for example, plant and operation efficiencies, energy mix and other country/region-specific data to provide a basis for realistic and reasonable comparison.

"Front-End" Tool Methodology

To this end, as part of the current work, a novel "Front-end" tool has been developed for "normalisation" of industry-acquired data (in terms of energy requirements and associated CO₂ emissions), as input to the benchmark methodology. To optimise the value of the output from the analysis, the nature and type of information required from industry about its recycling operations include:

- Commodity/commodities recovered at the plant
- Tonnage of feedstock through the plant per day
- Typical percentage of commodity fraction(s) in the feedstock
- Total energy consumption through the plant per day
- Indication of fuel source
- Tonnage of finished recovered product
- For a multi-commodity process, the fraction of energy usage attributable to the recovery of each commodity fraction.



As part of this approach, a questionnaire (Annex) has been designed to assist BIR in the acquisition of information from members and other stakeholders, and a fractionation flowchart methodology, shown in the following schematic, has been developed to take account of the information sought in situations where operations are handling either single- or multi-commodity materials at a plant, or are operating globally. Although presented as a metal fractionation flowchart, the cascade methodology can equally well be applied to non-metal fractions, such as high-grade polymers, paper/card or other materials.

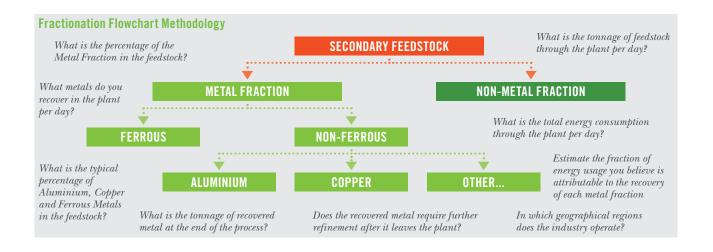
The "Front-end" tool, which uses a 2-step formula to determine the energy requirements and associated CO₂ emissions based on industry-acquired data, is presented below.

Step 1 – Data Acquisition and Fractionation Handling

This step involves acquisition of standardised data from industry through the use of a questionnaire (see p. 11) and the subsequent identification and apportionment of the energy usage, within the system boundary, for handling of each commodity fraction.

Based on the tonnage of feedstock (expressed as metric tonnes) identify the target fraction [A] and determine the percentage of each commodity fraction in [A] as $[B_{(1...n)}]$

Calculate the total energy consumption through the plant [C] per tonne of feedstock, taking account of the sum of energy use by fuel type $[\Sigma(C_1, C_n)]$ and,



Data sought from Industry

- Commodity/commodities recovered at the plant
- Tonnage of feedstock through the plant per day (t/d)
- Typical percentage of commodity fraction(s) in the feedstock (%)
- Total energy consumption through the plant per day per tonne of feedstock (MJ/t)
- Indication of "fuel" energy source (e.g. grid electricity, oil, diesel, coal etc)
- Tonnage of finished recovered product (t)
- For a multi-commodity process, the fraction of energy usage attributable to the recovery of each commodity fraction to deliver a product output

where applicable, the value for energy efficiency conversion [cf⁰] required to deliver the end product.

For each commodity fraction recovered $[\mathbf{D}_{(1...n)}]$, assign the energy usage $[\mathbf{E}_{(1...n)}]$:

Energy usage through plant in MJ per commodity fraction recovered:

$$E_{(1...n)} = D_{(1...n)} x (\Sigma(C_{1...}C_n) + cf^0)$$

Step 2 – Analysis and Normalisation of Energy Requirements

Express energy consumption in TJ as energy per 100,000t for each commodity product $[F_{(1...n)}]$ according to the following equation:

Energy in TJ/100,000t of product:

$$F_{(1...n)} = (E_{(1...n)} \div B_{(1...n)}) \div 10$$

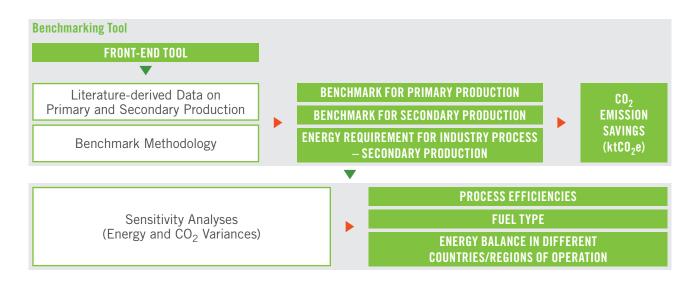
Express energy requirement for each commodity product $[\mathbf{F}_{(1...n)}]$ in $ktCO_2$ emission equivalent $[\mathbf{G}_{(1...n)}]$ per 100,000t of product based on:

- 1 Use of a conversion factor, [cf¹], to convert energy in TJ to energy in kWh (where 1TJ = 277,777.78kWh);
- 2 Grid electricity as the fuel source;
- 3 Use of a conversion factor, [cf²], to convert energy in kWh to CO₂ emissions in kgCO₂e; and
- 4 Use of a conversion factor, [cf³], to convert CO_2 emissions in kg CO_2 e to CO_2 emissions in kt CO_2 e (where 1kg CO_2 e = 0.000001kt CO_2 e)

CO₂ emissions per 100,000t of product:

$$G_{(1...n)} = F_{(1...n)} \times cf^1 \times cf^2 \times cf^3$$

For the purposes of comparison of the industry-acquired energy consumption and carbon footprint data with the literature-derived data, obtained in the 2008 study, for both primary and secondary production, the choice of life cycle boundaries is important to avoid the complications associated with differences in mining and beneficiation of ores and in the collection and transport of scrap to a recycling process, and is set as: (i) conversion of ore concentrate to metal in primary production, and (ii) from scrap and other secondary materials delivered to a recycling process and converted to metal in secondary production.



Benchmarking for BIR-Nominated Commodities

The commodities of interest to BIR in the current study are aluminium, copper, ferrous and paper. The production of each of these materials is set out below in the context of the processes involved, the energy consumption in these processes and the resulting carbon dioxide emissions. For the purpose of comparison, only those processes within the system boundaries are considered.

Aluminium

Primary and Secondary Aluminium Production

In 2014, the tonnages of primary and secondary aluminium produced were approximately 53Mt with about one third of aluminium demand satisfied from secondary production. According to the International Aluminium Institute (IAI), it has been reported that by 2020 metal demand is projected to have increased to around 97Mt (with around 31Mt recycled from scrap).

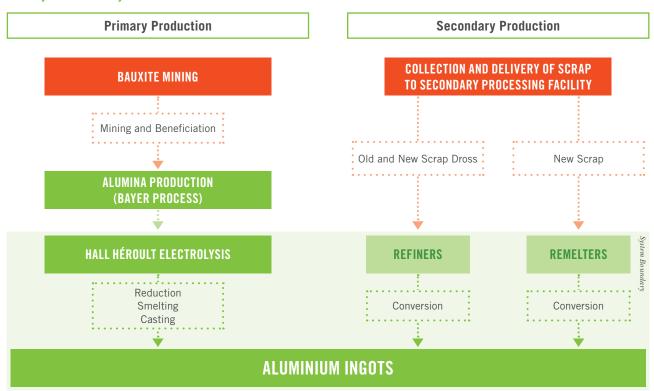
In the Bayer process, the bauxite ore is treated by alkaline digestion to beneficiate the ore. Although

the red mud produced in this process is a waste, which has major environmental impacts, comparison between primary and secondary aluminium production applying the benchmark methodology starts at the point of delivery of the alumina concentrate to the processing plant.

Primary production of aluminium from the ore concentrate is achieved by an electrolytic process in molten solution. The Hall Héroult process consists of electrolysis in molten alumina containing molten cryolite (Na₃AlF₆) to lower the melting point of the mixture from 2050°C for the ore concentrate to about 960°C.

The electrolysis cell consists of a carbon-lined reactor which acts as a cathode, with carbon anodes submerged in the molten electrolyte. In the electrolysis process, the aluminium produced is denser than the molten electrolyte and is deposited at the bottom of the cell, from where it is cast into ingots. At the anodes, the anodic reaction is the conversion of oxygen in the cell to carbon dioxide by reaction with the carbon of the anodes, and it is this reaction that is the major contributor to carbon

Primary and Secondary Production of Aluminium



dioxide emissions from the aluminium production process. The process results in the production of between 2% and 4% dross.

All secondary aluminium arisings are treated by refiners or remelters. Remelters accept only new scrap metal or efficiently sorted old scrap whose composition is known. Refiners, on the other hand, can work with all types of scrap as their process includes refinement of the metal to remove unwanted impurities. In both processes, the molten aluminium undergoes oxidation at the surface which has to be skimmed off as a dross. In Europe, about 2.5% of the feedstock aluminium in the refining process is converted to dross.

Process Energy Consumption and Carbon Dioxide Emissions

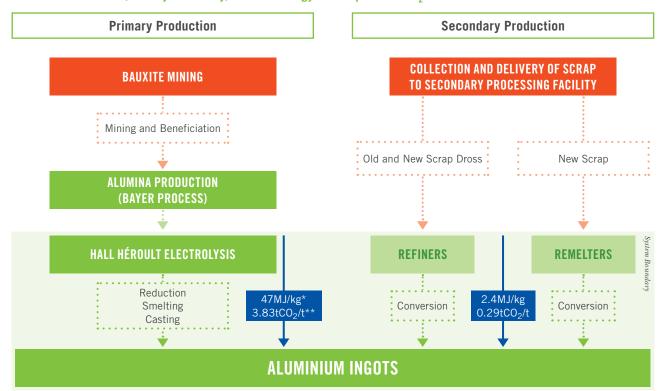
The gross energy requirement for primary aluminium production, via the Bayer Hall Héroult route, has been estimated at 120MJ/kg based on using hydroelectricity with 80% energy conversion efficiency [Norgate, CSIRO Minerals]. As alternatives to hydroelectricity, use of black coal for electricity

generation with an efficiency of 35% or natural gas with an efficiency of 54% would give gross energy estimates of approximately 211 and 150MJ/kg respectively.

The electricity consumption in the Hall Héroult process is the most energy-demanding aspect of primary production of aluminium. The energy requirements reported in the literature for the Hall Héroult process alone (i.e. for conversion of treated ore to metal) vary depending on the type of fuel used, with the values for electricity, for example, shown in the following table.

Energy Requirements – Hall Héroult Process only		
Source	MJ/kg Al	Notes
Schwarz	47	Electricity benchmark
IAI	54	Electricity average
Norgate	66	Electricity max.
Norgate	46	Electricity benchmark
IAI	69	Electricity max.

Aluminium Production (Primary/Secondary): Process Energy Consumption and CO₂ Emission



Source: For Primary Production: Benchmark Energy Value. *Schwarz: Carbon Footprint Value. **Choate and Green.

In the same way, the literature data on the carbon footprint for primary production of aluminium following the Bayer Hall Héroult route and for the Hall Héroult process alone vary on the type of fuel used as well as the conditions of the electrolysis reaction. For the Hall Héroult process alone, the carbon footprint values along with observations and assumptions made by the authors are shown in the following table.

Carbon Footprint – Hall Héroult Process only			
Source	Carbon Footprint (tCO ₂ /t AI)	Observations	
Norgate	7.2	Drained cathodes, Inert anodes, Low-temperature electrolytes, Natural gas 54%	
Norgate	4.6	Drained cathodes, Inert anodes, Low-temperature electrolytes, Hydroelectricity 89%	
IAI	7.7	Average IAI	
Choate and Green	3.83	US Average (typical)	

For the purpose of comparison of the energy requirements and associated carbon emissions for primary aluminium production with data for secondary aluminium production, an electricity benchmark figure of $47 \, \text{MJ/kg}$ and a CO_2 emission value of $3.83 \, \text{tCO}_2$ /t are used.

It has been reported that the production of one tonne of aluminium from scrap requires only 12% of the energy required for primary production. Energy savings of between 90 and 95% have also been reported for secondary aluminium production compared with primary production, starting with mining the ore and not with the as-received concentrate.

The energy requirement to recycle aluminium has been calculated at between 6 and 10MJ/kg assuming efficiencies of 60-80% in the recycling process.

Energy requirement data for secondary aluminium production from scrap are reported in the table

below as: mean values for melting and casting; and benchmark values for melting and casting. The corresponding carbon footprint data have been calculated on the basis of these energy requirement data using a carbon emission factor equivalent for the UK, for example, of 0.42857-0.47990ktCO₂e/kWh. According to the IAI, in 2009, the aluminium industry itself was reported as responsible for around 1% of the man-made greenhouse gas emissions, 40% of which result from the aluminium production process itself (direct emissions) with 60% resulting from electricity power generation (indirect emissions).

Secondary Production of Aluminium				
	Energy R	Requirement	CO ₂ E	missions
Source	Mean (MJ/kg)	Benchmark (MJ/kg)	Mean (tCO ₂ /t)	Benchmark (tCO ₂ /t AI)
Remelting	4.5	2.1	0.54	0.25
Casting	0.5	0.3	0.06	0.04
Total	5.0	2.4	0.60	0.29

Using the benchmark data for primary and secondary aluminium production from delivered ore concentrate and scrap respectively, the energy requirements for the production of 100,000 tonnes of aluminium are:

Energy requirement for primary production:	4700TJ
Energy requirement for secondary production:	240TJ

Using the energy data, the carbon footprints for primary and secondary production of aluminium on the same basis are:

Carbon footprint for primary production:	383ktC0 ₂
Carbon footprint for secondary production:	29ktC0 ₂

The benchmark energy and carbon footprint savings between primary and secondary aluminium production are 4460TJ/100,000t in energy and 354ktCO₂e/100,000t in CO₂ emissions, respectively.

Copper

Primary and Secondary Copper Production

According to the US Geological Survey, world copper production in 2013 was 17.9Mt. Copper recovered from scrap as a percentage of total copper produced is reported by the Copper Development Association in 2013 to be 35% for 2010 and 2011.

The major approach to primary copper production is the pyrometallurgical route from copper sulfide ores that have been concentrated usually by flotation to give the concentrate used in the pyrometallurgical process. A very small percentage of primary copper is recovered from copper ores hydrometallurgically.

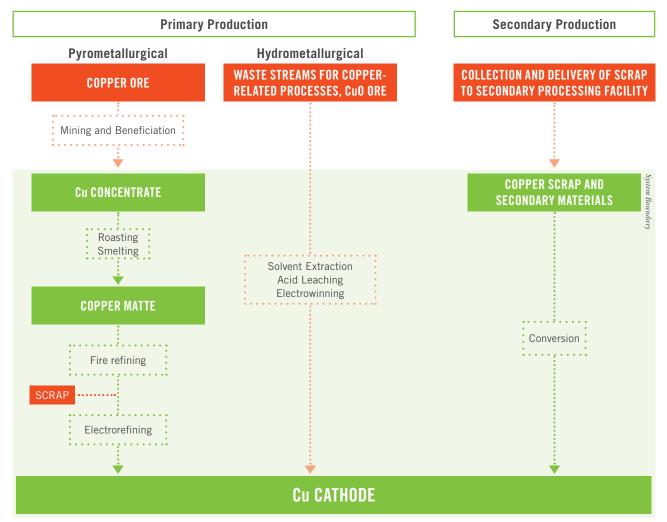
In the pyrometallurgical process, the concentrates are roasted to produce a copper matte which contains between 30 and 50% copper. The matte is

reduced to copper metal in a converter process, and the final product is generally purified by dissolving the copper metal obtained in sulfuric acid and recovering high-purity copper from this solution by electrowinning.

The hydrometallurgical route involves leaching of the copper oxide ore with sulfuric acid to produce a solution from which copper metal can be recovered on the cathodes of an electrowinning process.

Secondary copper can be produced from scrap and other copper-containing materials by pyrometallurgical and hydrometallurgical processes that are similar to those used in primary metal production.

Primary and Secondary Production of Copper



Process Energy Consumption and Carbon Dioxide Emissions

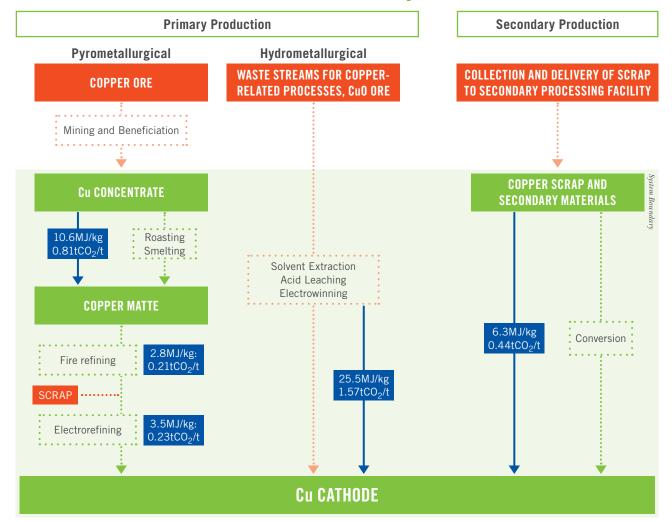
There are literature reports suggesting that the energy requirement for secondary copper production is between 35 and 85% of that for primary production. The International Copper Association (Copper Alliance) endorses the higher figure, commenting further that this equates to a saving of 40Mt of $\rm CO_2$ annually and the equivalent of 100,000 GWh of electricity.

The data for energy required for primary copper production via pyrometallurgical and hydrometallurgical routes are given in the following schematic, which also shows the point in the energy requirement diagram at which scrap copper would enter the pyrometallurgical process. These are the data on which comparisons between primary and secondary production have to be based.

The benchmark energy requirements for the production of cathode copper metal from primary copper ore concentrate, by pyrometallurgy, from soluble copper ores, by hydrometallurgy, and from scrap and secondary sources are shown in the above schematic and in the following table.

Copper Recovery Method	Energy Requirement (MJ/kg Cu)	Carbon Footprint (tCO ₂ /t Cu)
Pyrometallurgy from ore concentrate	16.9	1.25
Hydrometallurgy from oxide ores	25.5	1.57
Secondary production from scrap	6.3	0.44

Copper Production (Primary/Secondary): Process Energy Consumption and CO₂ Emissions



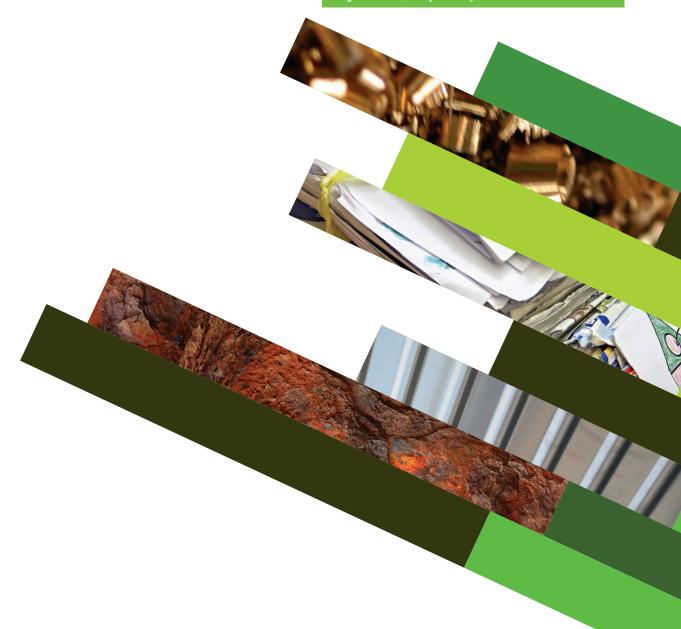
Using the benchmark data for primary and secondary copper production from delivered ore concentrate and scrap respectively, taken from the BIR 2008 report, the energy requirements for the production of 100,000 tonnes of copper are:

Energy requirement for pyrometallurgical primary production: 1690TJ
Energy requirement for hydrometallurgical primary production: 2550TJ
Energy requirement for secondary production: 630TJ

Using the energy data, the carbon footprints for primary and secondary production of copper on the same basis are:

Carbon footprint for pyrometallurgical primar production:	y 125ktCO ₂
Carbon footprint for hydrometallurgical prima production:	157ktC0 ₂
Carbon footprint for secondary production:	44ktC0 ₂

The benchmark energy and carbon footprint savings between primary (via the pyrometallurgical route) and secondary copper production are 1060TJ/100,000t in energy and $81ktCO_2e/100,000t$ in CO_2 emissions, respectively.



Ferrous

Primary and Secondary Ferrous Production

In 2013, world production of steel was reported by BIR and others as 1,607Mt with around 580Mt of scrap produced in the same year.

A schematic representation of iron recovery and steel manufacture is given in this section.

There are four main routes used for the production of steel, namely: blast furnace/basic oxygen furnace (BF-BOF); electric arc furnace (EAF); direct reduction (DR); and smelting reduction (SR).

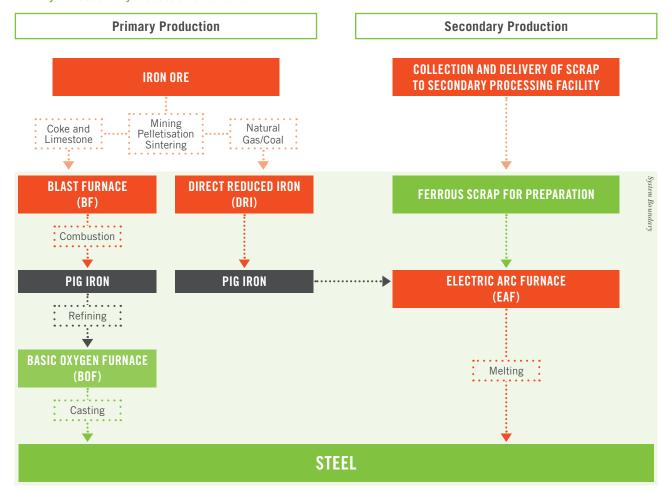
The BF-BOF route is the most complex and involves the reduction of iron oxide ore with carbon in the furnace.

Liquid iron produced in the blast furnace is referred to as pig iron and contains about 4% carbon. The

amount of carbon has to be reduced to less than 1% for use in steelmaking, and this reduction is achieved in a basic oxygen furnace (BOF) in which carbon reacts with oxygen to give carbon dioxide. The oxidation reaction is exothermic and produces enough energy to produce a melt. Scrap or ore is introduced at this stage to cool the mix and maintain the temperature at approximately 1600-1650°C. Blast furnaces consume about 60% of the overall energy demand of a steelworks, followed by rolling mills (25%), sinter plants (about 9%) and coke ovens (about 7%).

Direct reduction involves the production of primary iron from iron ores to deliver a direct reduced iron (DRI) product from the reaction between ores and a reducing gas in the reactor. The DRI product is mainly used as a feedstock in an electric arc

Primary and Secondary Production of Ferrous



furnace. The main advantage of this process is that the use of coke as a reductant is not required, thus avoiding the heavy burden on emissions resulting from coke production and use.

The electric arc furnace process involves the melting of DRI using the temperature generated by an electric arc formed between the electrode and the scrap metal, producing an energy of about 35MJ/s which is sufficient to raise the temperature to 1600°C. Depending on the quality of product required, the output of the EAF might need further treatment by secondary metallurgical and casting processes.

Smelting reduction (SR) is a current development that involves a combination of ore reduction and smelting in one reactor, without the use of coke. The product is liquid pig iron which can be treated and refined in the same way as pig iron from the blast furnace.

In secondary ferrous production, electric arc furnaces are used to produce steel from scrap using the same process as that described for the use of DRI as feedstock. Production of steel from scrap has been reported to consume considerably less energy compared to production of steel from iron ores.

Process Energy Consumption and Carbon Dioxide Emissions

The production of primary steel is more energy intensive than the production of secondary steel due to the chemical energy required to reduce iron ore to iron using reducing agents.

The literature values for the energy requirements and carbon footprints for the production of steel by different routes are in the following tables.

Energy Requirements for Steel Production from Ore Concentrate via BF-BOF Route

Source	Energy requirement (MJ/kg steel)
Ertem and Gurgen	16.58
Price et al	15.6
Phylipsen et al	15.47
Sakamoto	13.4
Mean (SD)	15.3 (1.3)

Carbon Footprint for Steel Production from Ore Concentrate via BF-BOF route

Source	Carbon footprint (tCO ₂ /t steel)
Norgate	2.3
Orth et al	2.23
Sakamoto	2.15
Orth et al	2.14
Das and Kandpal	2.12
Gielen and Moriguchi	2
Hu et al	1.97
Orth et al	1.82
Orth et al	1.69
Wang et al	1.32
Mean (SD)	1.97 (0.30)

Energy Requirements for Steel Production via DRI + EAF Combined Route

Source	Energy requirement (MJ/kg steel)	Note*
Das and Kandpal	36.9	Coal (India)
Das and Kandpal	24	Gas (India)
Price et al	19.2	80% DRI+20% scrap
Benchmark	19.2	

^{*} Noted assumptions made on the energy source used

Carbon Footprint for Steel Production via DRI + EAF Combined Route

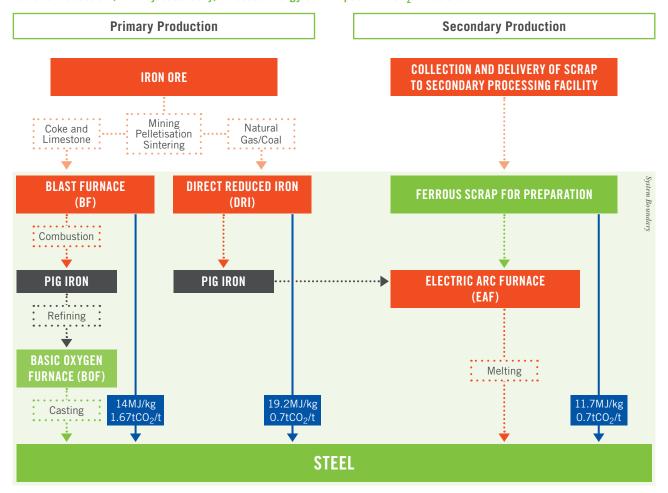
Carbon footprint (tCO ₂ /t steel)	Note
3.31	Coal (India)
1.74	Coal + Circofer
1.57	Gas
1.46	Gas + Circofer
0.7	Gas
1.76 (0.96)	
0.7	
	footprint (tCO ₂ /t steel) 3.31 1.74 1.57 1.46 0.7 1.76 (0.96)

 $^{^{\}ast}$ Noted assumptions made on the energy source used

Energy Requirements for Steel Production from Scrap via EAF Route				
Source	Energy requirement (MJ/kg steel)			
Das and Kandpal	14.4			
Hu et al	11.8			
Hu et al	11.2			
Sakamoto et al	9.4			
Mean (SD)	11.7 (2.1)			

Carbon Footprint for Steel Production from Scrap via EAF Route			
Source	Carbon footprint (tCO ₂ /t steel)		
Das and Kandpal	1.18		
Wang et al	0.64		
Hu et al	0.59		
Sakamoto et al	0.56		
Hu et al 0.54			
Mean (SD) 0.70 (0.27)			

Ferrous Production (Primary/Secondary): Process Energy Consumption and ${\rm CO_2}$ Emissions



The benchmark energy requirements for the production of steel from primary ore concentrate by the BF-BOF route, by the DRI + EAF route and from scrap and secondary sources via the EAF route are shown in the above schematic and in the following table.

Steel Recovery Method	Energy requirement (MJ/kg steel)*	Carbon Footprint (tCO ₂ /t steel)*
BF/BOF route	14	1.67
DRI + EAF route	19.2	0.7
EAF route	11.7	0.7

^{*}As reported and used in the 2008 study

Using the benchmark data for primary and secondary steel production from delivered ore concentrate and scrap respectively, the energy requirements for the production of 100,000 tonnes of steel are:

Energy requirement for primary production (BF-BOF route):	1400TJ
Energy requirement for primary production (DRI + EAF route):	1920TJ
Energy requirement for secondary production (EAF route):	1170TJ

Using the energy data, the carbon footprints for primary and secondary production of steel on the same basis are:

Carbon footprint for primary production (BF-BOF route):	167ktC0 ₂
Carbon footprint for primary production (DRI + EAF route):	70ktC0 ₂
Carbon footprint for secondary production (EAF route):	70ktC0 ₂

The benchmark energy and carbon footprint savings between primary (via the BF-BOF route) and secondary ferrous production are 230TJ/100,000t in energy and 97ktCO $_2$ e/100,000t in CO $_2$ emissions, respectively.



Paper

Primary and Secondary Paper Production

In 2012, global paper production amounted to 400Mt, according to BIR and other sources, with Asia accounting for 45% (182Mt), by far the largest paper producer. Europe (104Mt) and North America (85Mt) are also significant producers. Approximately 230Mt of recovered paper is collected worldwide from secondary sources, equivalent to 57% of primary production volumes.

Comparison of the primary and secondary papermaking industry is complicated for the following reasons:

- Recycled pulp and virgin pulp are often combined before manufacture;
- Paper can be recycled only 3-6 times before it degrades;
- The paper product from recycling will be of a lower quality than from primary sources;
- Some types of paper can be made only from 100% virgin pulp;
- Recycled pulp cannot be used alone some primary pulp is always required;
- Paper is produced from a renewable resource;
- As a waste, paper contains energy that can be recovered by incineration;
- Primary production removes trees and therefore reduces CO₂ uptake by the trees;
- Most literature compares disposal options rather than production options.

In the primary manufacture of paper, trees must be harvested, debarked, chipped at the sawmill and pulped with water. Pulping can be conducted by adding chemicals or by mechanical beating which will break down the lignin in wood and allow the pulp to form. Chemical pulping is expensive because the paper yield from wood is very low, but the paper produced is strong. Mechanical pulping is much cheaper despite the considerable use of electrical energy because it leads to a high yield of paper product from the wood, although this paper is much weaker. More water is introduced to the pulp before chemicals and dyes are added prior to the refining, screening and cleaning of the pulp which is then used in paper manufacture.

Waste paper from various sources is sorted, shredded, pulped with water and cleaned to remove impurities such as wire, plastic, paper clips and staples that may be in the mix. A de-inking cell cleans the pulp, removing ink and sticky substances. The pulp is fed into a blend chest where chemicals and dyes are added that will influence the character and appearance of the final product. The pulp is refined using a mechanical abrasive and bruising action before being screened, cleaned to remove any dirt or grit, and used in the manufacture of paper.

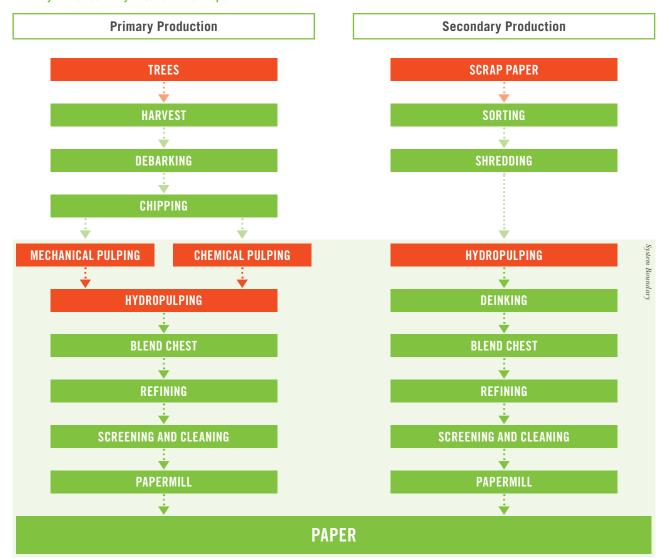
Process Energy Consumption and Carbon Dioxide Emissions

The global average energy requirements for the production of 1 tonne of paper has been reported as 10.8GJ of thermal energy and 4.5GJ of electrical energy, and it has been claimed that primary production requires 40% more energy than secondary production, but more fossil fuels are required to make secondary paper. It has also been reported that to produce paper from wood and then recycle it back into paper requires 22-53GJ/t, excluding transportation.

Schenk has compared energy requirements for chemically and mechanically processed paper manufacturing methods, using varying amounts of recyclate in the feedstock. For the use of 100% virgin pulp, the energy requirement is 12GJ/t for chemical processing and 28GJ/t for mechanical processing. If recycled pulp is added to the process, the energy consumption will be increased in the chemical process but decreased in the mechanical process.

A detailed report by The Paper Task Force in 2002 gave information on primary and secondary paper production. The types of paper considered were newsprint, corrugated, office paper and paperboard. The scope of the assessment was broad, including all activities involved from tree felling and waste paper collection to landfill consequences and incinerator ash disposal. For ease of comparison

Primary and Secondary Production of Paper



and simplification of analysis, it was considered that in the recycling scheme the pulp was 100% waste paper. In order to prepare paper of good quality, however, it would not be possible to use 100% pulp from recycled sources. Data obtained from the literature on energy use and direct ${\rm CO_2}$ emissions are in the following table.

Paper Recovery Method	Energy Requirement (MJ/kg paper)	Carbon Footprint (tCO ₂ /t paper)
Primary	35.2	0.0017
Secondary	18.8	0.0014

The data in the table below, which are for paper manufacturing steps only (i.e. excluding tree harvesting and transport, and waste paper collection and sorting), show that the total energy requirement for the recycling process is always less than the total energy for paper produced from virgin sources (see graph p. 26).

The benchmark figures for the energy requirement and carbon footprint for the manufacture of newsprint from primary and secondary sources are given in the following table.

Using the benchmark data for primary and secondary paper production from virgin pulp and scrap respectively, the energy requirements for the production of 100,000 tonnes of paper are:

Energy requirement for primary production:	3520TJ
Energy requirement for secondary production:	1880TJ

Using the energy data, the carbon footprints for primary and secondary production of paper on the same basis are:

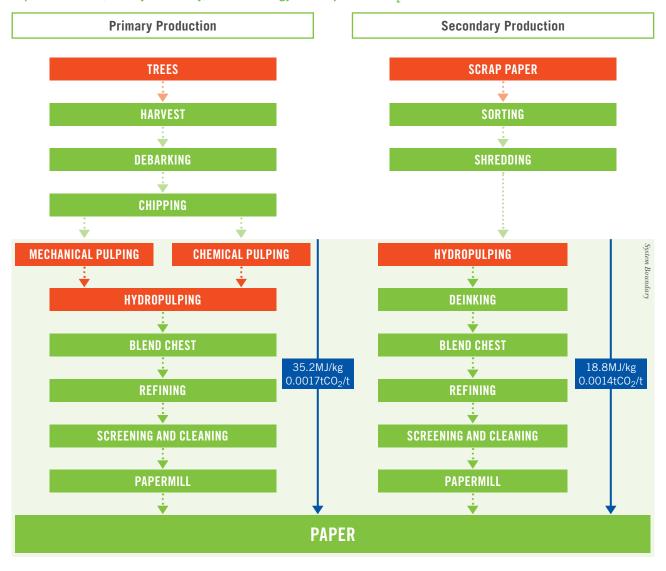
Carbon footprint for primary production:	170ktC0 ₂
Carbon footprint for secondary production:	140ktC0 ₂

The benchmark energy and carbon footprint savings between primary and secondary paper production are 1640TJ/100,000t in energy and 30ktCO₂e/100,000t in CO₂ emissions, respectively.

Energy Use and CO ₂ Emissions for Primary and Secondary Paper Production						
	UNITS	HAUTO Name de la	Corrugated	045	Paperboard	
	OM112	Newsprint	containers	Office paper	CUK ¹	SBS ²
		V	irgin Manufacture			
Total Energy	MJ/kg paper managed	39	28	40	28	41
Purchased Energy	MJ/kg paper managed	36	15	19	14	19
Fossil Fuel Energy	MJ/kg paper managed	26	13	14	12	14
		Re	cycled Manufactur	e		
Total Energy	MJ/kg paper managed	21	19	21	17	16
Purchased Energy	MJ/kg paper managed	21	19	21	17	16
Fossil Fuel Energy	MJ/kg paper managed	16	16	16	13	13
GHG Emissions — Whole system minus waste management and material recovery						
Virgin	CO ₂ eq t/t paper	0.0023	0.0014	0.0014	0.0009	0.0027
Recycle	CO ₂ eq t/t paper	0.0013	0.0013	0.0017	0.0013	0.0014

¹ CUK – Coated Unbleached Kraft

² SBS – Solid Bleached Sulfate



Summary

Using the benchmark data (from the 2008 study, derived from the literature) for primary and secondary production for the four BIR-nominated commodities of relevance to the current study – metals from delivered ore concentrate and scrap respectively, and paper from virgin pulp and scrap respectively - the energy requirements for the production of 100,000 tonnes of each commodity are shown in the following table.

Material	Primary (TJ)	Secondary (TJ)	Savings (TJ/100,000t)
Aluminium	4700	240	4460
Copper	1690	630	1060
Ferrous	1400	1170	230
Paper	3520	1880	1640

Using these energy data, the carbon footprints for primary and secondary commodity production on the same basis are shown in the following table.

Material	Primary (ktCO ₂ e)	Secondary (ktCO ₂ e)	Savings (ktCO ₂ e/100,000t)	% Savings (CO ₂ e)
Aluminium	383	29	354	92
Copper	125	44	81	65
Ferrous	167	70	97	58
Paper	0.17	0.14	0.03	18



Sensitivity Analysis and Impact of Externalities on Benchmark Data Interpretation

The benchmark energy and carbon emissions data for primary production calculated in the 2008 report represented the most efficient production processes available with the lowest energy consumption in situations where the best possible energy mixes are used anywhere in the world. It is important, however, to take account of operations that depart from the ideal benchmark conditions. Furthermore, in the current study, the effects of variations in energy conversion factors and energy source/fuel mix on the benchmark data are calculated to show the extent of their influence.

Sensitivity Analysis of Benchmark Data

To account for any such deviation, sensitivity analyses can be carried out on any of the input data in order to show how differences in process parameters would be reflected in the overall energy saving and carbon footprint results. Sensitivity analysis can be applied to the benchmark data and used for deviations from benchmark conditions including, for example:

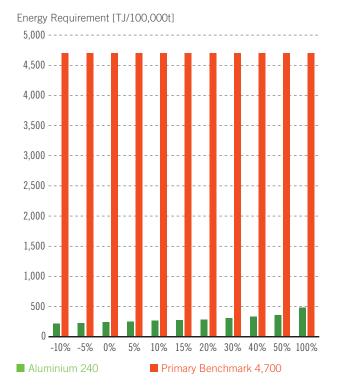
- Secondary production energy requirement data compared with the primary benchmark;
- Primary production energy requirement data from the benchmark;
- Primary production energy requirement data compared with secondary benchmark;
- Carbon footprint data for secondary recovery compared with the primary benchmark; and
- Carbon footprint data for primary production from the primary benchmark.

Both the primary and secondary benchmark figures will change with any plant inefficiencies compared to the benchmark conditions. The effect of these variations are shown numerically for an overall percentage deviation from the benchmark values (0% in the tables over the range -10% (i.e. more efficient) to +100% (less efficient)).

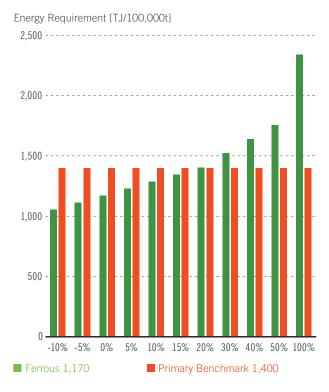
The variations are expressed for secondary energy requirement data vs. primary benchmark data in the table below, and diagrammatically in the charts on p. 28 for each of the BIR-nominated commodities.

Material	Primary	Sensitivity Analysis of Secondary Energy Data nary Secondary (expressed as TJ/100,000t)											
			-10%	-5%	0%	5%	10%	15%	20%	30%	40%	50%	100%
Aluminium	4700	240	216	228	240	252	264	276	288	312	336	360	480
Copper	1690	630	567	599	630	662	693	725	756	819	882	945	1260
Ferrous	1400	1170	1053	1112	1170	1229	1287	1346	1404	1521	1638	1755	2340
Paper	3520	1880	1692	1786	1880	1974	2068	2162	2256	2444	2632	2820	3760

Aluminium



Ferrous



The data show that, for aluminium and copper, even if a given process deviates by 100% from the benchmark, an energy saving would still be

Copper



Paper



predicted. However, for ferrous and paper, deviations of less than 100% would result in the prediction of an energy balance in favour of primary production.

By way of example, for the metals studied in the current work, the data in the following tables show:

- 1 How variations would arise in carbon footprint data for secondary production if a given process deviated from the primary benchmark data. The sensitivity analysis is calculated across the same range; for all the metals studied, deviations by plus 100% from the benchmark still lead to the prediction of carbon dioxide savings from the secondary process;
- 2 The variation in energy requirement and in CO₂ emissions with deviations from the primary benchmark data that would have to be compared with the energy requirement and the carbon footprint for secondary production in given situations.

Sensitivity analysis can also be applied to consider variations in energy and carbon footprint by country arising from differences in energy conversion factors and sources of energy and fuel/energy balance.

Metals Studied in	Primary	Secondary	Sensitivity Analysis of Secondary CO ₂ Data ary (expressed as ktCO ₂ e/100,000t)										
Current Work			-10%	-5%	0%	5%	10%	15%	20%	30%	40%	50%	100%
Aluminium	383	29	26	28	29	30	32	33	35	38	41	44	58
Copper	125	44	40	42	44	46	48	51	53	57	62	66	88
Ferrous	167	70	63	67	70	74	77	81	84	91	98	105	140

Metals Studied in Primary		Sensitivity Analysis of Primary Energy Data (expressed as TJ/100,000t)										
Current Work	-10%	-5%	0%	5%	10%	15%	20%	30%	40%	50%	100%	
Aluminium	4700	4230	4465	4700	4935	5170	5405	5640	6110	6580	7580	9400
Copper	1690	1521	1606	1690	1775	1859	1944	2028	2197	2366	2535	3380
Ferrous	1400	1260	1330	1400	1470	1540	1610	1680	1820	1960	2100	2800

Metals Studied in Current Work	Primary					vity Anal pressed						
Guilent Work		-10%	-5%	0%	5%	10%	15%	20%	30%	40%	50%	100%
Aluminium	383	345	364	383	402	421	440	460	498	536	575	766
Copper	125	113	119	125	131	138	144	150	163	175	188	250
Ferrous	167	150	159	167	175	184	192	200	217	234	251	334

Impact of Externalities on Benchmark Data

In this section the effects of variations in energy conversion factors and energy source/fuel mix on the benchmark data are calculated to show the extent of their influence.

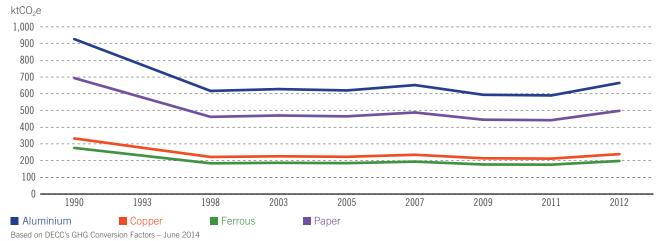
Energy Conversion Factors

Energy conversion factors, reported annually by country and region, in general, change year on

year. The UK Grid electricity generated (excluding imports), for example, has changed from a value of $0.70991 kgCO_2e/kWh$ in 1990 to the latest reported value of $0.50935 kgCO_2e/kWh$ in 2012 and calculated here to show the variation in the primary benchmark energy data (tabled and shown graphically on p. 30) for each of the BIR-nominated commodities over the last two decades.

	Energy Conversion	Aluminium (ktCO ₂ e)	Copper (ktCO ₂ e)	Ferrous (ktCO ₂ e)	Paper (ktCO ₂ e)				
Year	Factor kgCO ₂ e/kWh	Primary Benchmark Energy Data (TJ)							
		4700	1690	1400	3520				
1990	0.70991	927	333	276	694				
1993	0.59098	772	277	230	578				
1998	0.47226	617	222	184	462				
2003	0.48084	628	226	187	470				
2005	0.47515	620	223	185	465				
2007	0.49956	652	235	194	488				
2009	0.45501	594	214	177	445				
2011	0.45192	590	212	176	442				
2012	0.50935	665	239	198	498				

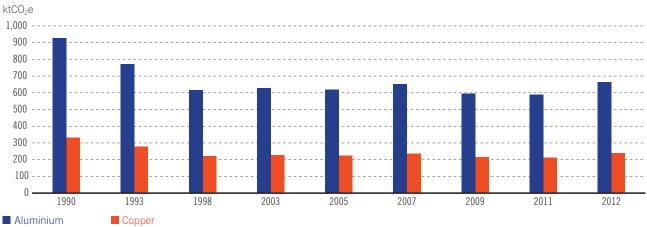
$Effect of Grid \ Electricity \ Emission \ Factor \ (in \ ktCO_2e/kWh) \ on \ Conversion \ of \ Primary \ Benchmark \ Energy \ Data \ to \ CO_2 \ Emissions \ CO_2 \ Emission \$



Using copper as an example commodity, it is noted from the calculated values for the energy requirement benchmark of 1690MJ, in the table above, that there is a variance of $121ktCO_2e$ in CO_2 emissions between a maximum of $333ktCO_2e$

recorded in 1990 to a minimum of $212ktCO_2e$ in 2011. By comparison, for aluminium there is a variance of $337ktCO_2e$ in CO_2 emissions over the same time period. A profile for both commodities is shown below.

Effect of Grid Electricity Emission Factor (in $ktCO_2e/kWh$) on Conversion of Primary Benchmark Energy Data to CO_2 Emissions: Profile for Aluminium and Copper



Based on DECC's GHG Conversion Factors – June 2014

Energy Source/Fuel Mix

One obvious potential variation in energy use between countries or regions depends upon the nature of the energy source, ranging from efficient hydroelectric production of electricity to the use of low-grade coals.

Many primary and secondary production processes for metals rely, for example, on electricity as a source of energy, and the following table lists different sources of electricity (GWh) extracted from data contained in DECC's guidance document on GHG Conversion Factors (June 2014).

Source	Electricity (GWh)
Gas	176,748
Coal	126,699
Nuclear	52,486
Hydro	9,257
Biomass	8,090
Wind	7,097
Oil	6,101
Waste	2,871
Solar PV	17

Source: Based on DECC's GHG Conversion Factors - June 2014

The associated ${\rm CO}_2$ emission factors for different fuel types, taken from the Carbon Trust (drawn from DECC's GHG Conversion Factors for 2011, published

in June 2013 (DECC, 2013)), are shown in the following table, with the most recently reported value of $0.50935 kgCO_2e/kWh$ for grid electricity from DECC (2014) – the value used in the current study – included for reference.

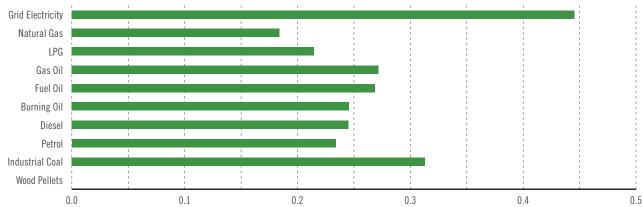
Fuel Type	Energy Conversion Factor (kgCO ₂ e/kWh)
Grid Electricity ^a	0.50935
Grid Electricity	0.44548
Coal	0.31304
Gas Oil	0.27176
Fuel Oil	0.26876
Burning Oil	0.24555
Diesel (contains biofuel content)	0.24512
Petrol (contains biofuel content)	0.23394
LPG	0.21452
Natural Gas	0.18404
Wood Pellets	0

Source: Carbon Trust drawn from DECC's GHG Conversion Factors – June 2013.
^a Based on DECC's GHG Conversion Factors - June 2014.

The energy conversion factors are quoted as total "direct" $kgCO_2e$ per unit of fuel, as representing the emissions at the point of use of a fuel or at the point of generation for electricity. A profile of the comparative energy conversion factors is shown in the graph below.

Comparative Energy Conversion Factors per Fuel Type

Energy Conversion Factor kgCO₂e/kWh



Source: Carbon Trust drawn from DECC's GHG Conversion Factors – June 2013

Influence of Energy Conversion Factors & Source/Fuel Mix on Benchmark Data

According to the International Energy Agency (IEA), in 2012 global CO₂ emissions were 31.7 GtCO₂, representing a 1.2% year-on-year increase in emissions, about half the average annual growth rate since 2000, and four percentage points less than in 2010, the year of initial recovery after the financial crisis. The IEA provides information on the electricity emission factors (expressed as CO₂ emissions per kWh) worldwide between 1990 and 2012, as well as averaged for 2009-2011. These data are used here in calculations to take account of variations in electricity-based energy conversion factors in different EU countries and across different world regions; and energy conversion data for any fuel or mixed fuel use. Using copper as an example, the results of the calculations are reported in tables and profiles below.

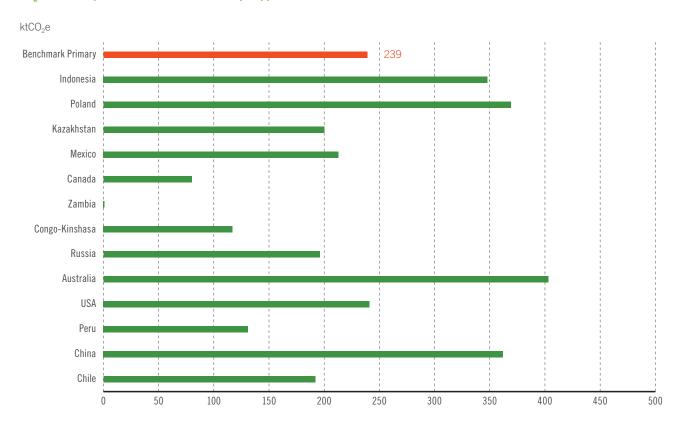
The estimated primary production of copper in 2013 was 17.9M tonnes, with almost 90% mined in the largest producing countries, as listed in the table. Corresponding CO_2 emissions per kWh from electricity generation are included.

Largest Copper Mine Producers and Production in 2013								
Country	Copper Mine Production (Mt) ^a	CO ₂ Emissions kgCO ₂ e/kWh ^b						
Chile	5.70	0.408						
China	1.65	0.771						
Peru	1.30	0.280						
USA	1.22	0.514						
Australia	0.99	0.859						
Russia	0.93	0.417						
Congo-Kinshasa (incl DRC)	0.90	0.250						
Zambia	0.83	0.003						
Canada	0.63	0.171						
Mexico	0.48	0.454						
Kazakhstan	0.44	0.427						
Poland	0.43	0.787						
Indonesia	0.38	0.741						
Total	15.88	6.082						

^a U.S. Geological Survey, Mineral Commodity Summaries, February 2014;

^b IEA Statistics 2013 CO₂ Emissions from Fuel Combustion

CO₂ Emissions per 100,000 tonnes of Primary Copper Production in Selected Countries



The production of 100,000 tonnes of primary copper following the pyrometallurgical route can be compared for different countries on the basis of electricity emission factors.

To do this, it is assumed that all countries initially have the same specific energy requirement (based on the benchmark primary and secondary data of 1690TJ per 100,000 tonnes and 630TJ per 100,000 tonnes respectively), and takes no account of variation in process efficiencies at country level.

Assuming that the energy requirement can be based on electricity emission factors, the profile on p. 32 shows the calculated emissions of CO₂ per 100,000 tonnes of primary copper production from ore concentrate compared with the benchmark value of 239ktCO₂, using the energy conversion factor of 0.50935kgCO₂e/kWh.

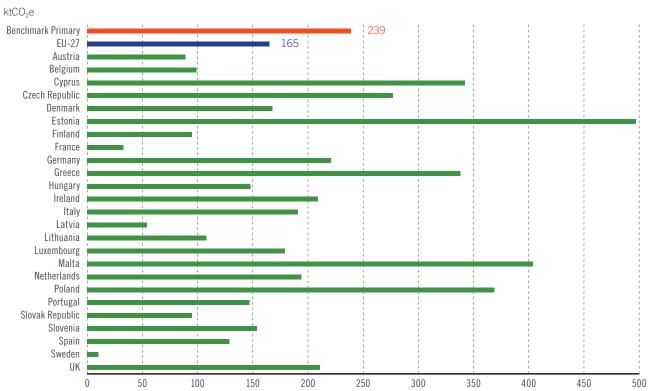
The data are based on the most recent reported figures from the IEA (averaged for the period 2009-2011) published in 2013 for ${\rm CO_2}$ emissions per kWh from electricity generation, by country, and have been applied to the benchmark primary energy requirement of 1690TJ, normalised to provide direct

comparison. From this comparison profile, four producer countries, for example, exceed by more than 10% the benchmark value of 239ktCO₂, with primary production in Australia (at 403ktCO₂) some 69% higher than the benchmark.

A similar profile (not shown) is obtained for CO₂ emissions for secondary copper production from scrap, normalising the data against the benchmark secondary energy requirement of 630TJ and again taking no account of changes in process efficiencies at country level. On this basis, comparison with the secondary benchmark value of 89ktCO₂ indicates a CO₂ emission level for secondary production in Australia that exceeds the benchmark value by 150ktCO₂.

The profile below shows a sensitivity analysis that takes account of the differences in the electricity-based energy conversion factors for different countries across the EU and shows how the $\rm CO_2$ emissions would change (for copper) if the conversion factors for other countries were used rather than the conversion value of 0.50935kg $\rm CO_2$ e/kWh used in the study.

Variation in Primary Carbon Footprint Data for Copper (based on Conversion Factors for EU-27 Countries)



Source: IEA CO₂ Emissions from Fuel Combustion Highlights 2013

Hypothetically, looking at the same profile across Europe, with the EU-27 average at 165ktCO_2 , there are six countries that would exceed the benchmark value of 239ktCO_2 , with Estonia generating an emission level of 497ktCO_2 , some 108% higher. Sweden, on the other hand, records the lowest level of emissions at 10ktCO_2 , 96% lower than the benchmark value.

It can be seen from the profile that the calculated UK value of 211ktCO_2 is less than the benchmark of 239ktCO_2 . The UK value is calculated using the IEA's most recent data for CO_2 emissions per kWh from electricity generation for all countries (published in 2013), a figure which is close to that reported in the UK DECC GHG Conversion Factors for 2011 (DECC, 2013), but less than the value used to calculate the benchmark equivalent which is taken from the UK DECC GHG Conversion Factors for 2012 (DECC, 2014), 0.50935kgCO $_2$ e/kWh.

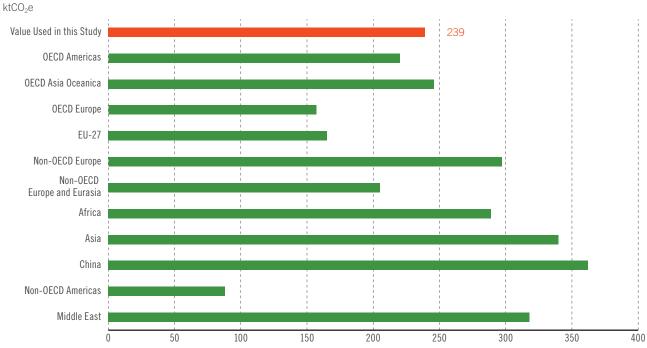
Interpreting the IEA figures for electricity emission factors (expressed as $\rm CO_2$ emissions per kWh), by region, for 2009-2011, and normalising the data against the primary benchmark of 239ktCO₂, the profile below shows that five would exceed this level by more than 10%, with China, a major producing country generating an emission level of 362ktCO₂,

some 50% higher, with the Non-OECD Americas region, on the other hand, recording the lowest comparative level of emissions at $88ktCO_2$, or 63% lower than the benchmark value.

Regional differences in contributions to global emissions, however, do conceal even larger differences among individual countries as shown in the following table.

Region	Calculated ktCO ₂ e
OECD Europe	157
Estonia	497
Non-OECD Europe and Eurasia	297
Kosovo	576
Africa	289
Botswana	839
Non-OECD Americas	88
Cuba	475
Benchmark	239

Variation in Primary Carbon Footprint Data for Copper (based on Conversion Factors for World Regions)

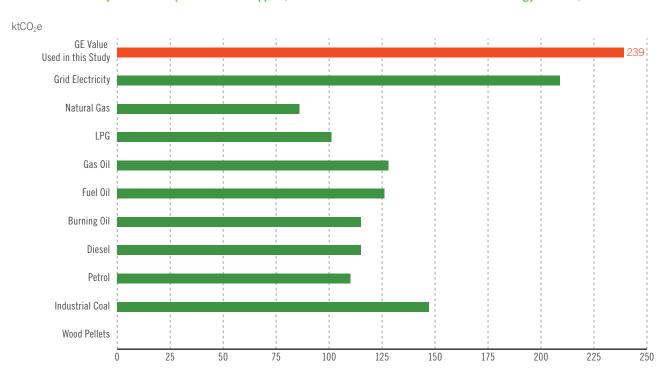


Source: IEA CO₂ Emissions from Fuel Combustion Highlights 2013

Using the comparative most recent energy conversion factors for electricity generation, by fuel type, cited earlier in this section, the data can be normalised against the primary benchmark of 239ktCO₂. The resulting profile below shows the

impact of a change in energy conversion factor for grid electricity [from $0.44548 \text{kgCO}_2 \text{e/kWh}$ (reported for 2011) to $0.50935 \text{kgCO}_2 \text{e/kWh}$ (reported for 2012)], on CO_2 emissions – 209ktCO_2 and 239ktCO_2 respectively.

Variation in Primary Carbon Footprint Data for Copper (based on Conversion Factors for Different Energy Sources)



Summary

The benchmark energy and carbon emissions data for primary production represent the most efficient production processes available with the lowest energy consumption in situations where the best possible energy mixes are used anywhere in the world. To take account of operations that depart from the ideal benchmark conditions for the production of the BIR-nominated commodities, sensitivity analyses have been carried out on the data to show how differences in process parameters would be reflected in the overall energy saving and carbon footprint results.

The impact of variations in energy conversion factors, energy source, fuel mix and country emission levels have been described. Although illustrated for copper as the example commodity, using its benchmark primary and secondary data, the impact of these and other variations such as process efficiencies can be studied for any commodity using benchmark or real industry data.

Analysis and Interpretation of Industry-Acquired Data for Single- and Mixed-Stream Processes

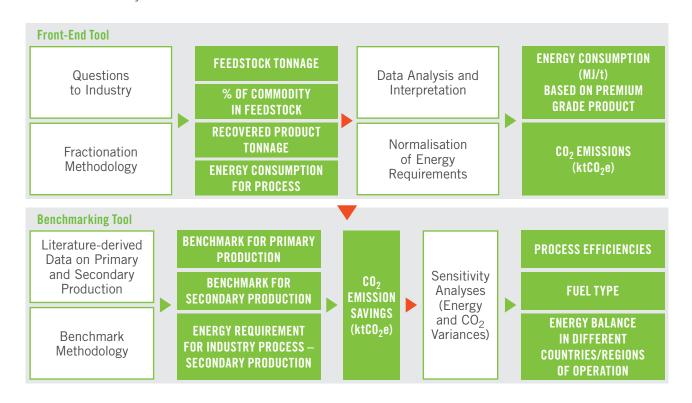
Introduction

To realise the full potential and value of applying the benchmark methodology, developed as part of the 2008 BIR study, relies on industry engagement such that the energy requirements and carbon footprints can be determined based on industry-acquired data. In this section, data provided from industry have allowed calculations of energy requirements and carbon dioxide emissions for both single-stream processes for each of the BIR-nominated commodities, and separately for mixed-metal streams containing all three target metals. As agreed with BIR, the sources of these industry data remain anonymous and unattributed.

Using the data provided by industry, the energy requirement for the recovery of 100,000 tonnes of each commodity is calculated based on the methodologies developed and applied in this study (illustrated in the schematic below) and compared with the benchmark values (essentially the best achievable) for its primary and secondary production derived from the desk-based research carried out in the 2008 BIR study.

Conversion of these energy requirements to the corresponding CO₂ emissions for the industry-acquired data used, for illustrative purposes, the most recently reported value of the energy conversion factor for UK Grid electricity generated (excluding imports), namely, 0.50935kgCO₂e/kWh (taken from the UK DECC GHG Conversion Factors for 2012, published in June 2014 (DECC, 2014)). Sensitivity analyses on these data are carried out to illustrate the effects of deviations from benchmark conditions. To permit direct in-region comparison to be made, the 2008 benchmark data have been recalculated using the 0.50935kgCO₂e/kWh energy conversion factor.

For the purposes of interpretation, values obtained directly from industry are referred to as "industry-acquired" but any values reported based on calculations in this work are referred to as "industry-derived".



Single-Stream Process

For single-stream processes, industry data have been provided that have allowed the determination of energy requirements and CO_2 emissions (using the "Front-end" tool) for the recovery of a fully-refined product for each commodity. Moreover, sensitivity analyses can be used on the industry-acquired data for each commodity, for individual process plants and regions, to illustrate how differences in operating conditions and process parameters can influence the overall energy savings and carbon footprint results.

Aluminium

The data on energy and carbon dioxide emissions savings are presented for aluminium delivered as a fully-refined product. The calculated energy requirement for the recovery of 100,000 tonnes of aluminium is compared with the benchmark values for primary and secondary production, as shown below.

Energy requirement for primary production:	4700TJ
Energy requirement for secondary production:	240TJ
Energy requirement for the process	
(industry-derived):	266TJ

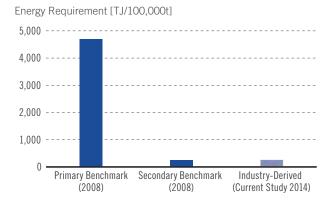
These data are shown diagrammatically for the commodity to illustrate the comparisons (see below).

Using the energy data and applying the energy conversion factor of 0.50935kgCO₂e/kWh, the carbon footprints (CO₂ emissions) for primary and secondary processes of aluminium, on the same basis, are:

Carbon footprint for primary production:	665ktC0 ₂
Carbon footprint for secondary production:	34ktC0 ₂
Carbon footprint for the process	
(industry-derived):	38ktC0 ₂

The benchmark energy and carbon footprint savings achieved by industry against the primary benchmark for aluminium production in a region where the conversion factor, 0.50935kgCO₂e/kWh, is appropriate, would be 4434TJ/100,000t in energy and 627ktCO₂e/100,000t in CO₂ emissions, respectively.

Aluminium



Copper

The data on energy and carbon dioxide emissions savings are presented for copper delivered as a fully-refined product. The calculated energy requirement for the recovery of 100,000 tonnes of copper is compared with the benchmark values for primary and secondary production, as shown below:

Energy requirement for primary production:	1690TJ
Energy requirement for secondary production:	630TJ
Energy requirement for the process	
(industry-derived):	657TJ

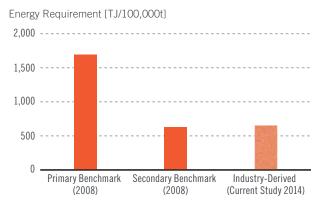
These data are shown diagrammatically for the commodity to illustrate the comparisons (see below).

Using the energy data and applying the energy conversion factor of 0.50935kgCO₂e/kWh, the carbon footprints (CO₂ emissions) for primary and secondary processes of copper, on the same basis, are:

Carbon footprint for primary production:	239ktC0 ₂
Carbon footprint for secondary production:	89ktC0 ₂
Carbon footprint for the process	
(industry-derived):	93ktC0 ₂

The benchmark energy and carbon footprint savings achieved by industry against the primary benchmark for copper production via the pyrometallurgical route, in a region where the conversion factor, 0.50935kgCO₂e/kWh, is appropriate, are 1033TJ/100,000t in energy and 146ktCO₂e/100,000t in CO₂ emissions, respectively.

Copper



Ferrous

The data on energy and carbon dioxide emissions savings are presented for ferrous delivered as a fully-refined product. The calculated energy requirement for the recovery of 100,000 tonnes of ferrous is compared with the benchmark values for primary and secondary production, as shown below:

Energy requirement for primary production:	1400TJ
Energy requirement for secondary production:	1170TJ
Energy requirement for the process	
(industry-derived):	1194TJ

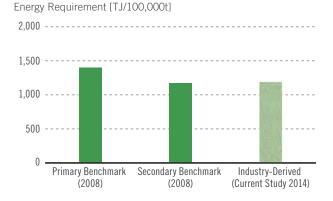
These data are shown diagrammatically for the commodity to illustrate the comparisons (see below).

Using the energy data and applying the energy conversion factor of 0.50935kgCO₂e/kWh, the carbon footprints (CO₂ emissions) for primary and secondary processes of ferrous, on the same basis, are:

Carbon footprint for primary production:	198ktC0 ₂
Carbon footprint for secondary production:	166ktC0 ₂
Carbon footprint for the process	
(industry-derived):	169ktCO ₂

The benchmark energy and carbon footprint savings achieved by industry against the primary benchmark for ferrous production via the BF-BOF route in a region where the conversion factor, 0.50935kgCO₂e/kWh, is appropriate, are 206TJ/100,000t in energy and 29ktCO₂e/100,000t in CO₂ emissions, respectively.

Ferrous



Paper

The data on energy and carbon dioxide emissions savings are presented for paper delivered as a fully-refined product. The calculated energy requirement for the recovery of 100,000 tonnes of paper is compared with the benchmark values for primary and secondary production, as shown below:

Energy requirement for primary production:	3520TJ
Energy requirement for secondary production:	1880TJ
Energy requirement for the process	
(industry-derived):	1541TJ

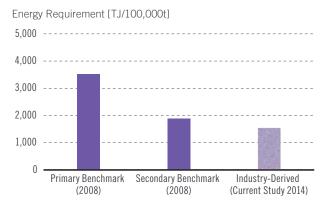
These data are shown diagrammatically for the commodity to illustrate the comparisons (see below).

Using the energy data and applying the energy conversion factor of 0.50935kgCO₂e/kWh, the carbon footprints (CO₂ emissions) for primary and secondary processes of paper, on the same basis, are:

Carbon footprint for primary production:	498ktC0 ₂
Carbon footprint for secondary production:	266ktC0 ₂
Carbon footprint for the process	
(industry-derived):	218ktC0 ₂

The benchmark energy and carbon footprint savings achieved by industry against the primary benchmark for paper production in a region where the conversion factor, 0.50935kgCO₂e/kWh, is appropriate, are 1979TJ/100,000t in energy and 280ktCO₂e/100,000t in CO₂ emissions, respectively.

Paper



Sensitivity Analysis of Industry-Derived Data

Sensitivity analysis can be applied to industryacquired data for individual process plants and regions to take account of differences in operating conditions and illustrate the effects of deviations from benchmark conditions.

The data in the tables below show how overall variations, expressed as a percentage deviation from the industry-derived data, would affect the energy savings (and CO_2 emissions) achieved in specific plant processes. It should be noted for aluminium, copper and paper that even if the savings in energy consumption are doubled, there are still substantial savings in energy consumption (and associated CO_2 emissions using the energy conversion factor of $0.50935 \mathrm{kgCO}_2 \mathrm{e/kWh}$ (DECC, 2014)) arising from the recovery of these secondary materials.

In the case of ferrous materials, however, overall inefficiencies of 20% or more (assuming that the primary benchmark situation can be achieved in every case) do eliminate the savings across the production plant, although it is likely that if the differences in the energy required to deliver ore to the plant and secondary material to the plant are considered, this would still favour secondary production.

For each of the BIR-nominated commodities, aluminium, copper, ferrous and paper, the sensitivity analyses, expressed as a general percentage variation of industry-derived secondary data for the material (from the current study) and the secondary benchmark data (from the 2008 study), are compared with the single benchmark value for primary production (also derived in the 2008 study).

Material	Primary	Industry-Derived Secondary Energy Requirement Data Secondary (expressed as TJ/100,000t) Industry-											
		Derived	-10%	-5%	0%	5%	10%	15%	20%	30%	40%	50%	100%
Aluminium	4700	266	239	253	266	279	293	306	319	346	372	399	532
Copper	1690	657	591	624	657	690	723	756	788	854	920	986	1314
Ferrous	1400	1194	1075	1134	1194	1254	1313	1373	1433	1552	1672	1791	2388
Paper	3520	1541	1387	1464	1541	1618	1695	1772	1849	2003	2157	2312	3082

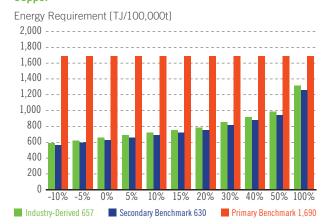
Material	Primary												
		Derived*	-10%	-5%	0%	5%	10%	15%	20%	30%	40%	50%	100%
Aluminium	665	38	34	36	38	40	41	43	45	49	53	57	75
Copper	239	93	84	88	93	98	102	107	112	121	130	139	186
Ferrous	198	169	152	161	169	177	186	194	203	220	237	253	338
Paper	498	218	196	207	218	229	240	251	262	283	305	327	436

^{*} Based on DECC's GHG Conversion Factors - June 2014

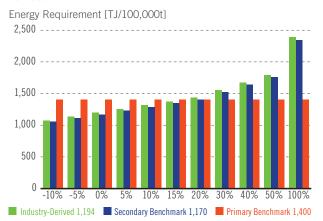
Aluminium



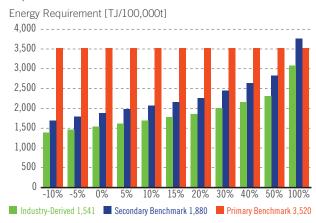
Copper



Ferrous



Paper



Copper													
Primary	Secondary	Secondary Secondary Recovery Data – Carbon Footprint dary Industry- (expressed as ktCO₂e/100,000t)											
		Derived	-10%	-5%	0%	5%	10%	15%	20%	30%	40%	50%	100%
239	89*		80	85	89	93	98	102	107	116	125	134	178
239		93*	84	88	93	98	102	107	112	121	130	139	186

^{*} Based on DECC's GHG Conversion Factors - June 2014

The table above shows how variations would arise in carbon footprint data for secondary production, drawn from literature data, and separately, acquired from industry if a given process deviated from the primary benchmark data. The sensitivity analysis, shown here for copper as an example, is calculated across the same range; for the metal, deviations by plus 100% from the benchmark still lead to the prediction of carbon dioxide savings from both secondary processes.

Copper: Sensitivity Analysis



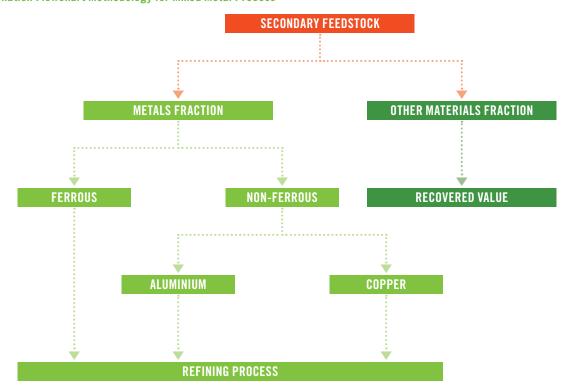
Mixed-Stream Process

In the same way, for mixed-stream processes, industry data have been provided that have allowed the determination of energy requirements and ${\rm CO_2}$ emissions for the recovery of a "saleable" product for further refinement to a fully-refined product, according to the fractionation flowchart methodology below.

In situations where data have been provided for a mixed waste stream comprising both metal and non-metal fractions, applying the "Front-end" tool it is possible to determine the energy requirements for the recovery of the commodity metal fractions assuming that: (i) the non-metal "other materials" fraction is producing a return and not going to waste and therefore has no negative energy cost, and (ii) some or all of that non-metal fraction does go to waste and therefore the energy can be allocated to recovery of each of the commodity materials.

Using the data provided by industry and applying the 2-step formula of the "Front-end" tool developed as part of this study, the energy usage through the plant per commodity fraction recovered $(\mathbf{E}_{(1...n)})$ from the mixed-stream process is calculated.

Fractionation Flowchart Methodology for Mixed Metal Process



Step 1

Energy usage through plant in MJ per commodity fraction recovered: $E_{(1...n)} = D_{(1...n)} \; x \; (\Sigma(C_{1...}C_n) + cf^0)$

Following step 1, these data can be expressed in terms of energy consumption ($\mathbf{F}_{(1...n)}$) and corresponding CO_2 emissions ($\mathbf{G}_{(1...n)}$) for the recovery of 100,000 tonnes of each commodity product, using the formulae set out in step 2:

Step 2

Energy in TJ/100,000t of product:
$$\begin{split} F_{(1...n)} &= (E_{(1...n)} \div B_{(1...n)}) \div 10 \\ \text{CO}_2 \text{ emissions per 100,000t of product:} \\ G_{(1...n)} &= F_{(1...n)} \text{ x cf}^1 \text{ x cf}^2 \text{ x cf}^3 \end{split}$$

The energy requirements for the recovery of aluminium, copper and ferrous metals from a mixed-stream process are calculated and shown in the following table. Applying the energy conversion factor of $0.50935 kgCO_2e/kWh$ (reference cf^1 in the equation above), the associated CO_2 emissions for aluminium, copper and ferrous have been determined assuming all of the energy is used to produce saleable products of recovered aluminium, recovered copper and recovered ferrous.

Material	Energy Requirements for Secondary Production (TJ/100,000t)	CO ₂ Emissions for Secondary Production* (ktCO ₂ e/100,000t)
Aluminium	27	33
Copper	27	34
Ferrous	24	30

^{*} Based on DECC's GHG Conversion Factors - June 2014

The energy values shown above could increase to 266, 657 and 1194 for aluminium, copper and ferrous respectively (in line with the calculated industry-derived values summarised for these metals reported on pages 37 and 38) if the operation was treating materials to a stage that needed then to go to refinement (e.g. to a smelter) and the refinement could be carried out at the same site. But if the material had to be transported offsite, the energy requirements could increase beyond these values to take account of the additional energy costs associated with transport.

Whilst it is unlikely that industry will not be able to recover value from the residual non-targeted material, consideration can be given to the reallocation of the energy, attributed to this fraction, to recover more of the target material(s). To take account of situations where less than 100% recovery of useful material is achieved from

a mixed-stream process, sensitivity analysis on the data can be applied.

Use of the "Front-end" tool can be extended to allow calculation of the potential for redeploying the energy attributed to the non-target fraction (up to 100%) to recovery of further value from the commodity material(s). Considering the mixed stream of aluminium, ferrous and copper, with energy requirements of 27, 27 and 24TJ per 100,000 tonnes respectively as recovered product, and other non-target material, the following formula permits calculation of the reallocated energy $[F^R_{(1...n)}]$ to one or more commodity materials. Using ferrous metal as an example, the reallocated energy, $F^R_{(Fe)}$, for the ferrous commodity is calculated as:

Reallocated Energy in TJ/100,000t of product:
$$F^{\text{R}}_{(Fe)} = (E_{(Fe)} + (E_{(Fe)} \times D_{(0ther)}) \div B_{(Fe)}) \div 10$$

where

- 1 **D**_(Other) is the non-target waste fraction;
- 2 **B**_(Fe) is the percentage of the ferrous fraction in the target fraction of a mixed stream; and
- 3 **E**_(Fe) is the energy usage through the plant in MJ per ferrous fraction recovered.

The total allocation of the energy, both baseline and reallocated residual energy, for ferrous is shown in the table below.

The data show the extent to which additional energy can be channelled towards recovery of ferrous product with a maximum level of 30TJ per 100,000 tonnes assuming 25% of the energy is not associated with the recovery of the target commodity metals.

Total Allocation of Energy for Ferrous Metal (TJ/100,000t)											
Danilling Francis		0%	5%	10%	15%	20%	30%	40%	50%	90%	100%
Baseline E	Baseline Energy		25	26	28	29	31	34	36	46	48
Reallocation of Additional	100%	30	32	33	35	37					
Energy	75%	29	30	32	33	35					
	50%	27	29	30	31	33					
	25%	26	27	28	29	31					

Summary

Where operations are handling either single- or multi-commodity materials at a plant, use of the novel "Front-end" tool, developed in the work, allows determination of energy requirements and CO₂ emissions for the recovery of a fully-refined product for each commodity from a single-stream process; and recovery of a "saleable" product for further refinement to a fully-refined product from mixed-stream processes. Data used in these calculations were provided by industry, the sources of which remain anonymous and unattributed.

In mixed-stream processes from which value can be derived from more than one commodity stream, where account needs to be taken of the energy apportioned to each metal and non-metal fraction, fractionation and attribution of energy data in the process have been applied to determine the energy and CO₂ emissions for each recovered commodity product, and, where there is less than 100% recovery of useful material, the tool can be extended to attribute that energy back into further recovery of the commodity product(s).

Sensitivity analyses have been applied to the benchmark and industry-acquired data for each commodity, for individual process plants and regions, to take account of differences in operating conditions and process parameters and to show any such deviation would be reflected in the overall energy saving and carbon footprint results.

Furthermore, there are two ways in which the results from the current study can be compared with those from the 2008 study as set out below:

Regional Comparison: A comparison of the industryderived data (from secondary production) with the benchmark primary energy data from the 2008 study assuming that the primary production can be carried out in the same region as the secondary production.

The benchmark primary energy data from the 2008 study can be converted to CO_2 emissions data using the same conversion factor used for the industry-derived data which effectively gives a comparison between primary production and the industry-acquired secondary process energy and CO_2 emissions savings data assuming that primary production could be achieved in a region where this conversion factor is appropriate. These results, which are comparable to those derived in the literature-based 2008 study, can be extended to processes carried out in any region or country by use of the relevant conversion factor.

The energy requirements for single-stream processes for the recovery of aluminium, copper, ferrous or paper are compared with the benchmark values (essentially the best achievable) for primary production from the desk-based research carried out in the 2008 study, as shown in the table below. The energy savings achieved against the primary benchmark, using the industry-acquired data, are 4434, 1033, 206 and 1979 TJ/100,000t for each commodity respectively.

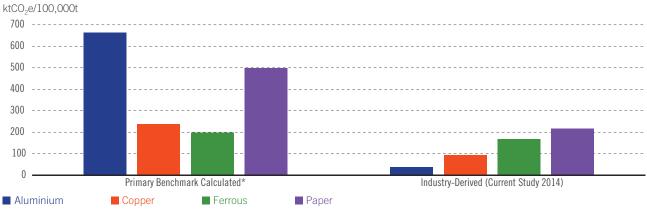
Using the energy data and applying the energy conversion factor, for the UK, of $0.50935 kgCO_2e/kWh$, the carbon footprints (CO_2 emissions) for primary and industry-derived processes for aluminium, copper, ferrous and paper are presented in the table and graph on p. 44. The corresponding savings in CO_2 emissions achieved against the primary benchmark, using the industry-acquired data, are 627, 146, 29 and $280ktCO_2e/100,000t$ for each commodity respectively.

Material	Energy Requ	Savings (achieved by industry	
	Primary Benchmark (TJ)	Industry-Derived (TJ)	against Primary Benchmark) (TJ/100,000t)
Aluminium	4700	266	4434
Copper	1690	657	1033
Ferrous	1400	1194	206
Paper	3520	1541	1979

Carbon Footprint (CO $_2$ emissions) for Primary and Industry-derived Processes					
Material	Calculated CO ₂ Emissions* from Primary Energy Benchmark (ktCO ₂ e/100,000t)	Industry-Derived CO ₂ Emissions* (ktCO ₂ e/100,000t)	Savings (achieved by industry against Primary Benchmark) (ktCO ₂ e/100,000t)		
Aluminium	665	38	627		
Copper	239	93	146		
Ferrous	198	169	29		
Paper	498	218	280		

^{*} Based on the same Conversion Factor (DECC's GHG Conversion Factors - June 2014)

Carbon Footprint (CO₂ emissions) for Primary and Industry-derived Processes



Based on DECC's GHG Conversion Factors – June 2014

Best Available Worldwide Comparison:

A comparison of the industry-derived data (from secondary production) with the benchmark primary data from the 2008 study representing the most efficient production processes available with the lowest energy consumption in situations where the best possible energy mixes are used anywhere in the world.

In the 2008 report, highlight calculations of total CO₂ savings arising from the benchmark primary data for annual worldwide secondary material production were made. In the current study, energy

savings obtained for the production of 100,000t of secondary aluminium, copper and ferrous metals, using the industry-acquired data, were compared with primary production benchmark data from the 2008 study and converted to CO_2 emissions savings for the most recent worldwide annual secondary production tonnages for each commodity metal studied in the current work, with total estimated savings in annual CO_2 emissions arising from the secondary production of the metals, shown in the table below, in comparison with primary production, of 572Mt.

Material	Energy Savings (achieved by industry against Primary Benchmark) (TJ/100,000t)	Annual Worldwide Secondary Production* (Mt)	Estimated Savings in Annual CO ₂ Emissions (Mt)	
Aluminium	4434	18	63.3	
Copper	1033	6	4.8	
Ferrous	206	580	503.9	
	in Annual CO ₂ Emissions Secondary Metals Studied [Current Study]		572.0	

^{*} Annual worldwide secondary production (Mt) as quoted in 2014 for Aluminium and in 2013 for Copper and Ferrous

Conclusion and Summary Highlights

Conclusion

The benefits of the 2008 research have been widely acknowledged in the recycling industry worldwide, and, for the recycling industries, the value of expressing environmental benefits in terms of $\rm CO_2$ emissions savings is becoming increasingly necessary. It is in this context that the current study is carried out:

"... to update the findings of the 2008 study, applying the methodology developed, to more recent data on four of the original commodities, namely, aluminium, copper, ferrous metals and paper".

The 2008 desk-based research involved a detailed review of available scientific and technical literature on seven metals - aluminium, copper, ferrous metals, lead, nickel, tin and zinc - and of paper. The study introduced the concept and application of a benchmark methodology to determine the best available and most justifiable energy use and carbon emissions data for primary production processes, and used best estimates from the literature of benchmark data for energy and carbon footprint calculations for both primary and secondary production. For primary production, the benchmark data represented the most efficient production processes available with the lowest energy consumption per tonne of metal produced in situations where the best possible energy mixes were used. The conversion factors used to express the primary production energy data as benchmark carbon emission data were also based on those for the best possible energy mixes. Benchmark data were, thus, defined as those data that represented material production situations that were achievable and gave values that were most acceptable and justifiable as the best achievable, acknowledging that these would not be achieved in all primary production processes. The calculations of benchmark values for secondary production in the 2008 study were similarly derived from literaturebased data.

The benchmark data were used to highlight the advantages (environmental impacts) of secondary production over primary production and were reported per 100,000 tonnes of material produced to provide a means of direct comparison between primary and secondary production and expressed

as CO₂ savings per 100,000 tonnes of production. To avoid complications associated with the early stages of the whole lifecycles of these materials, benchmark energy requirements and carbon footprints were taken from ore or raw material delivered at the production plant for primary material, and delivered at the secondary plant for secondary material. Sensitivity analyses were then developed and used to show how the benchmark results can be used to deal with variations in different production processes, for example, variations in efficiency, and fuel and energy balances

Building on the acknowledged benefits of the 2008 research, it is clear that realising the full potential and value of applying the benchmark methodology relies on industry engagement so that the energy requirements and carbon footprints are determined on industry-acquired data (i.e. "real" data). In the current study, use has been made of industryacquired data to obtain energy and carbon emission results from real situations. As with the 2008 study, sensitivity analyses have been used in this work to enable the data presented to be extended to other operations to take account of differences in, for example, plant and operation efficiencies, energy mix and other country/region-specific data to permit realistic and reasonable comparisons to be made in any situation with any set of variable factors.

As part of the current study, a novel "Front-end" tool has been developed for "normalisation" of industry-acquired data (in terms of energy requirements and associated CO₂ emissions), as input to the methodology used. To optimise the value of the output from the analysis, the nature and type of information required from industry about its recycling operations formed the basis of a questionnaire to industry to assist BIR in the acquisition of information from members and other stakeholders. A fractionation flowchart methodology was developed to take account of the information sought in situations where operations are handling either single- or multi-commodity materials at a plant, or are operating globally. The "Front-end" tool uses a two-step formula to determine the energy requirements and associated CO₂ emissions based on industry-acquired data.

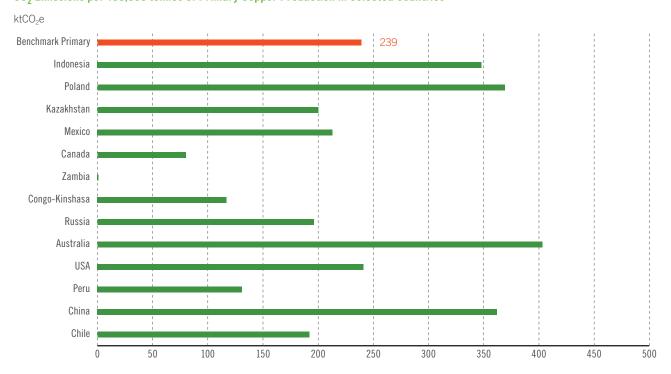
As part of the update of carbon emission data for the commodities of interest, using industry-derived data, an investigation of the impact of variations in energy conversion factors and energy source/fuel mixes on the benchmark data has been carried out. These factors not only change with region but also with time. For example, the energy conversion factor for UK Grid electricity generated changed from 0.70991kgCO₂e/kWh in 1990 to the latest reported value of 0.50935kgCO₂e/kWh in 2012.

Initial calculations in this study make use of the most recently published electricity conversion factors (taken from UK DECC GHG Conversion Factors (DECC, 2014)) to convert energy data to carbon emissions. To obtain a comparison of the results with the possible primary production for the same region (e.g. the UK), the electricity conversion factor of 0.50935kgCO₂e/kWh has also been applied to the 2008 benchmark data for primary production. All data and graphical and schematic representations being derived from these calculations are reported here for the first time.

Another potential variation in energy use between countries or regions depends upon the nature of the energy source, ranging from efficient hydroelectric production of electricity to the use of low-grade coals. To look at the influence of energy source/ fuel mix on benchmark data, copper has been used as an example. The production of 100,000 tonnes of primary copper following the pyrometallurgical route can be compared, for example, for different countries on the basis of electricity emission factors. To do this, it is assumed that all countries initially have the same specific energy requirement (based on the benchmark primary and secondary data of 1690TJ per 100,000 tonnes and 630TJ per 100,000 tonnes respectively), and takes no account of variation in process efficiencies at country level.

Assuming that the energy requirement can be based on electricity emission factors, the following profile shows the calculated emissions of $\rm CO_2$ per 100,000 tonnes of primary copper production from ore concentrate compared with the benchmark value of 239ktCO₂, using the energy conversion factor of 0.50935kgCO₂e/kWh.

CO₂ Emissions per 100,000 tonnes of Primary Copper Production in Selected Countries



The data are based on the most recent reported figures from the IEA (averaged for the period 2009-2011) published in 2013 for CO_2 emissions per kWh from electricity generation, by country, and have been applied to the benchmark primary energy requirement of 1690TJ, normalised to provide direct comparison.

Other profiles reported in the current work show how sensitivity analyses can take account of differences in the electricity-based energy conversion factors for different countries across the EU and global regions and show how the CO_2 emissions would change if the conversion factors for other countries were used rather than the conversion value of $0.50935 \mathrm{kgCO}_2\mathrm{e/kWh}$ used in the study.

Data provided from industry (which remain anonymised) have allowed calculations of energy requirements and carbon dioxide emissions for both single-stream processes for each of the BIR-nominated commodities, and separately, for mixed-metal streams containing all three target metals. Using these data, the energy requirement for the recovery of 100,000 tonnes of each commodity has been calculated based on the methodology developed and compared with the benchmark values (essentially the best achievable) for its primary and secondary production derived from the desk-based research carried out in the 2008 BIR study.

For the single-stream process, the industry data provided allow for the determination of energy requirements and CO₂ emissions for the recovery of a fully-refined product for each commodity. Sensitivity analyses have been applied to the benchmark and industry-acquired data for each commodity, for individual process plants and regions, to take account of differences in operating conditions and process parameters and to show any such deviation would be reflected in the overall energy saving and carbon footprint results. For mixed-stream processes, the industry data provided allow the determination of energy requirements and CO₂ emissions for the recovery of a "saleable" product for further refinement to a fully-refined product. The calculated energy requirements and CO₂ emissions for secondary production from this mixed stream for aluminium, copper and ferrous are 27, 27 and 24 TJ per 100,000 tonnes and 33, 34 and 30 ktCO₂e per 100,000 tonnes respectively. In situations where industry is handling mixed streams

from which value can be derived from more than one commodity stream, account needs to be taken of the energy apportioned to each metal and nonmetal fraction. In this regard, use has been made of the "Front-end" tool, developed in this work, to systematically attribute energy data in the process and determine the energy and $\rm CO_2$ emissions for each recovered commodity product. In situations where there is less than 100% recovery of useful material, the tool can be extended to attribute that energy back into further recovery of the commodity product(s).

Summary Highlights

The summary highlights from the current study are:

- 1 Realising the full potential and value of applying the benchmark methodology relies on industry engagement such that the energy requirements and carbon footprints are determined on industry-acquired data (i.e. "real" data). Of further benefit is the ability to apply sensitivity analyses on these data to take account of differences in, for example, plant and operation efficiencies, energy mix and other country/region-specific data to provide a basis for realistic and reasonable comparison;
- 2 Using the methodology developed and applied in this work, variations in energy conversion factors, energy source, fuel mix and country emission levels, and other variations such as process efficiencies, do impact the CO₂ emissions, as illustrated for copper, for any commodity, exploiting either benchmark or real industry data;
- 3 Where operations are handling either single- or multi-commodity materials at a plant, or are operating globally, use of the novel "Front-end" tool developed in the work allows determination of energy requirements and CO₂ emissions for the recovery of a fully-refined product for each commodity from a single-stream process; and recovery of a "saleable" product for further refinement to a fully-refined product from mixed-stream processes. In mixed-stream processes from which value can be derived from more than one commodity stream, fractionation and attribution of energy data in the process have been applied to determine the energy and

- CO₂ emissions for each recovered commodity product, and, where there is less than 100% recovery of useful material, the tool can be extended to attribute that energy back into further recovery of the commodity product(s);
- 4 Using the same conversion factor as that used for the industry-derived data for converting the benchmark primary energy data from the 2008 study to CO₂ emissions data allows comparison of the industry-derived data (from secondary production) with the benchmark primary energy data, assuming that primary production can be carried out in the same region where this conversion factor is appropriate [Regional Comparison].

The calculated energy savings achieved, against the primary benchmark, are 4434, 1033, 206 and 1979 TJ/100,000t for aluminium, copper, ferrous and paper respectively, with corresponding savings in CO₂ emissions (using the electricity conversion factor, for the UK, of 0.50935kgCO₂e/kWh) of 627, 146, 29 and 280ktCO₂e/100,000t for each commodity respectively. These results, which are comparable to those derived in the literature-based 2008 study, can be extended to processes carried out in any region or country by use of the relevant conversion factor; and

5 With the benchmark primary data from the 2008 study representing the most efficient production processes available with the lowest energy consumption in situations where the best possible energy mixes are used anywhere in the world, comparison of the industry-derived data with the benchmark primary data can be made using updated figures of total savings in annual CO₂ emissions for worldwide secondary production [Best Available Worldwide Comparison].

The BIR-nominated metals (aluminium, copper and ferrous metals) studied in the current work, represented 98% of the total estimated annual savings in CO₂ emissions (equivalent to 488.3Mt/100,000t) that would result from worldwide secondary production of the materials contained in the 2008 report. Using the most recent worldwide secondary production tonnages for these metals, the estimated savings in annual CO₂ emissions arising from the secondary production of aluminium, copper and ferrous metals, in comparison with primary production, are 572Mt.

Furthermore, the methodologies described in this report can be used to obtain energy and carbon emissions data with a view to calculating the potential CO_2 savings that can be achieved for:

- 1 Recycling operations of any BIR member for any of the nominated commodities described in this work – aluminium, copper, ferrous metals and paper;
- Other ferrous-based commodities [stainless steel and its alloys – containing chromium, molybdenum, nickel, titanium, tungsten and vanadium] of importance to BIR and its members; and
- 3 Any other material of interest to any BIR member, on a case-by-case basis, in the industry's secondary recovery processes.

Bibliography (Update on 2008 Report)

- Aluminium International Today Buyers' Directory, A Report on Environmental Benefits of Recycling – A Critical Review of the Data for Aluminium, 2010, 4-7
- Agrawal, A., and Sahu, K.K., *Problems, prospects and current trends of copper recycling in India: An overview*, Resources, Conservation and Recycling, 2010, 54, 401-416
- Aluminium production: an example calculation of the energy saved by recycling. Department of Materials Science and Metallurgy, University of Cambridge. Online accessed: 14 October 2014. www.doitpoms.ac.uk/tlplib/recycling-metals/aluminium_production.php
- Alvarado, S., Maldonado, P., and Jaques, I., Energy and environmental implications of copper production, Energy, 1999, 24, 307–316
- Amira International Limited, Copper Technology Roadmap, 2004, 1-50
- Andersen, J.P., and Hyman, B., Energy and material flow models for the U.S. steel industry, Energy, 2001, 26, 137-159
- Ayres, R.U., Ayres, L.W., and Rade, I., *The life cycle of copper, its co-products and by-products*, International Institute for Environment and Development and the World Business Council for Sustainable Development, 2002, 24, 1-210
- Balomenos, E., Panias, D., and Paspaliaris, I., Energy and Exergy Analysis of the Primary Aluminium Production Processes: A Review on Current and Future Sustainability, Mineral Processing and Extractive Metall. Rev., 2011, 32, 69-89
- Beer, J.D., Worrell, E., and Blok, K., Long-term energy efficiency Improvements in the paper and board industry, Energy, 1998, 23(1), 21-42
- Bhaktavatsalam, A.K., and Choudhury, R. Specific energy consumption in the steel industry, Energy, 1995, 20, 1247-1250
- Blanco, A.A., Negro, C., Monte, C., Fuente, E., and Tijero, J., The challenges of sustainable papermaking, Environmental Science & Technology, 2004, 38(21), 414A-420A
- Bloemhof-Ruwaard, J.M., Van Wassenhove, L.N., Gabel, H.L., and Weaver, P.M., *An environmental life cycle optimization model for the European pulp and paper industry*, Omega, International Journal of Management Science, 1996, 24(6), 615-629
- Blomberg, J. and Hellmer, S., Short-run demand and supply elasticities in the West European market for secondary aluminium, Resources Policy, 2000, 6, 39-50
- Boin, U.M.J., and Bertram, M., Melting Standardized Aluminium Scrap: A Mass Balance Model for Europe JOM, 2005, 26-33
- Boustead, I., and Hancock, G.F., *Energy and Materials Requirements of Primary Aluminium Production in the UK*,
 Resource, Recovery and Conservation, 1981, 5, 303-318
- Brogaard, L.K., Damgaard, A., Jensen, M.B., and Barlaz, M., Evaluation of Life Cycle Inventory Data for Recycling Systems, Resources, Conservation and Recycling, 2014, 87, 30-45
- Bruch, H., Gohike, D., Kruger, C., Reuter, M., Roepenack, I.V., Rombach, E., Rombach, G., and Winkler, P. LCI of copper production and processing, Metall., 1995, 49(4), 252-257
- Burdick, D., Energy and greenhouse gas impacts of mining and mineral processing operations, J Cleaner Production, 2010, 18(3), 266-275
- Magnaghi, G., Recovered Paper Market in 2012 BIR Global Facts and Figures, Bureau of International Recycling Paper Division, 2014, 1-24

- Carbon Trust, Conversion Factors Energy and Carbon Conversions 2013 Update, 2012, 1-10 www.carbontrust.com
- Chalmin, P., and Journet, C., World Markets for Recovered and Recycled Commodities: The End of the "Waste Era"..., Commissioned by Bureau of International Recycling, 2011, 1-52
- Chen, W-Q., and Graedel, T.E., *Dynamic analysis of aluminium stocks and flow in the United States: 1900-2009*, Ecological Economics, 2012, 81, 92-102
- Choate, W.T., and Green, J.A., *U.S. Energy requirements for aluminium production: Historical perspective, theoretical limits and new opportunities.* U.S. Department of Energy, Energy Efficiency and Renewable Energy, 2003, 12-24
- Choudhury, R., and Bhaktavatsalam, A.K., *Energy inefficiency of Indian steel industry scope for energy conservation*, Energy Conservation and Management, 1997, 38(2), 167-171
- Ciacci, L., Eckelman, M.J., Passarini, F., Chen, W-Q., Vassura, I., and Morselli, L., *Historical Evolution of Greenhouse Gas emissions from Aluminium Production at a Country Level*, J Cleaner Production, 2014, 1-10
- Ćirković, M., Trujić, V., and Bugarin, M., Synergy of energy resources of copper pyrometallurgy in RTB Bor-Serbia, Metall. Mater. Eng., 2014, 20(4), 261-273
- Confederation of European Paper Industries, *Paper Industry Statistics*, [Cited 11/01/15] Available from: www.cepi.org
- CRU Strategies, *Global Non-Ferrous Scrap Flows 2000-2011* with a focus on Aluminium and Copper, Commissioned by Bureau of International Recycling, 2013, 1-68
- Damgaard, A., Larsen, A.W., and Christensen, T.H., Recycling of metals: accounting of greenhouse gases and global warming contributions, Waste Management & Research, 2009, 27, 773-780
- Das, A., and Kandpal, T.C., *Iron and Steel Manufacturing Technologies in India: Estimation of CO₂ Emission*, Intl. J. of Energy Research, 1997, 21, 1187-1201
- Das, A., and Kandpal, T.C., Energy demand and associated ${\it CO_2}$ emissions for the Indian steel industry, Energy, 1998, 23(12), 1043-1050
- Das, S., Achieving Carbon Neutrality in the Global Aluminium Industry, JOM, 2012, 64(2), 285-290
- Das, S., Aluminium Recycling in a Carbon Constrained World: Observations and Opportunities, JOM, 2011, 63(8), 137-140
- Das, S.K., Long III, W.J., Hayden, W., Green, J.A.S., and Hunt, W.H.Jr., *Energy Implications of the Changing World of Aluminium Metal Supply*, JOM, 2004, 14-17
- DECC/Defra, 2013 Government GHG Conversion Factors for Company Reporting: Methodology Paper for Emission Factors, July 2013, 1-111
- DECC, 2014 Government GHG Conversion Factors for Company Reporting: Methodology Paper for Emission Factors, July 2014, 1-112
- Ding, N., Gao, F., Wang, Z., Gong, X., and Nie, Z., *Environment Impact Analysis of Primary Aluminium and Recycled Aluminium*, Procedia Engineering, 2012, 27, 465-475
- Dresher, W.H., How Hydrometallurgy and the SX/EW Process Made Copper the Green Metal, Copper Development Association Inc., 2001, http://copperalliance.org

- Du, J.D., Han, W.J., Peng, Y.H., and Gu, C.C., Potential for Reducing GHG emissions and Energy Consumption from Implementing the Aluminium Intensive Vehicle Fleet in China, Energy, 2010, 35, 4671-4678
- Dzioubinski, O., and Chipman, R., Trends in Consumption and Production: Selected Minerals, 1999, 1-18
- Eckelman, M.J., Ciacci, L., Kavlak, G., Nuss, P., Reck, B.K., and Graedel, T.E., *Life Cycle Carbon Benefits of Aerospace Alloy Recycling*, J Cleaner Production, 2014, 80, 38-45
- Emi, T., Changing paradigm of metal separation technology for steel production, Scandinavian Journal of Metallurgy, 2005, 34, 79–88
- Ertem, M., and Gürgen, S., *Energy balance analysis for Erdemir blast furnace number one*, Applied Thermal Engineering, 2006, 26, 1139-1148
- Eurofer The European Steel Association, Steel Production -Energy Efficiency Working Group Final Report, 2014, 1-55
- European Aluminium Association, Environmental Profile Report for the European Aluminium Industry - Life Cycle Inventory data for aluminium production and transformation processes in Europe April 2013 – Data for the Year 2010, 2013, 1-78
- European Aluminium Association, Environmental Profile Report for the European Aluminium Industry - Life Cycle Inventory data for aluminium production and transformation processes in Europe April 2008 – Data for Year 2005, 2008, 1-84
- European Aluminium Association, Global Aluminium Recycling: A Cornerstone of Sustainable Development, 2009, 1-36
- European Aluminium Association, *Aluminium Use in Europe:* Country Profiles 2007-2010, 2011, 1-27
- Farla, J., Blok, K., and Schipper, L., *Energy Efficiency Developments in the Pulp and Paper Industry*, Energy Policy, 1997, 25(7-9), 745-758
- Forestry Commission, *UK Wood Production and Trade 2013 Provisional Figures*, 2014, 1-24
- Frees, N., Crediting Aluminium Recycling in LCA by Demand or by Disposal, Intl J LCA, 2008, 13(3) 212-218, (DOI:http://dx.doi.org/10.1065/lca2007.06.348)
- Gaballah, I., and Kanari, N., *Recycling policy in the European Union*, JOM, 2001, 53(11), 24-27
- Gaudreault, C., Samson, J., and Stuart, P., Implications of choices and interpretation in LCA for multi-criteria process design: de-inked pulp capacity and cogeneration at a paper mill case study, J Cleaner Production, 2009, 17, 1535-1546
- Gaustad, G., Olivetti, E., and Kirchain, R., Improving Aluminium Recycling: A Survey of Sorting and Impurity Removal Technologies, Resources, Conservation and Recycling, 2012, 58, 79-87
- Gielen, D., and Moriguchi, Y., CO_2 in the iron and steel industry: An analysis of Japanese emission reduction potentials, Energy Policy, 2002, 30, 849-863
- Gilstad, G. *Life Cycle Assessment of Secondary Aluminium Refining*, NTNU-Trondheim Master of Energy and Environmental Engineering, June 2013
- Giurco, D., and Petrie, J.G., Strategies for reducing the carbon footprint of copper: New technologies, more recycling or demand management? Minerals Engineering, 2007, 20(9), 842–853

- Gloser, S., Soulier, M., and Tercero Espinoza, L.A., *Dynamic Analysis of Global Copper Flows. Global Stocks, Postconsumer Material Flows, Recycling Indicators, and Uncertainty Evaluation* Environ.Sci.Technol., 2013, 47, 6564-6572
- Gonzalez Palencia, J.C., Furubayashi, T., and Nakata, T., Analysis of CO₂ Emissions Reduction Potential in Secondary Production and Semi-fabrication of Non-Ferrous Metals, Energy Policy, 2013, 52, 328-341
- Graus, W., and Worrell, E., Methods for calculating CO₂ intensity of power generation and consumption: A global perspective, Energy Policy, 2011, 39, 613–627
- Green, J.A.S., (Editor) Aluminium Recycling and Processing for Energy Conservation and Sustainability, AMS International, 2007
- Guan, Y., Shao, C., Tian, X., and Ju, M., Carbon Footprint Attributed to Aluminium Substitution for Copper in the Chinese Indoor Air Conditioner Industry, J Cleaner Production, 2013, 51, 126-132
- Guo, Z.C., and Fu, Z.X., Current Situation of Energy Consumption and Measures taken for Energy Saving in the Iron and Steel Industry in China, Energy, 2010, 35, 4356-4360
- Hammond, G., and Jones, C., *Inventory of Carbon and Energy* (ICE) Version 1.6a, University of Bath/Carbon Trust/EPSRC, 2008, 1-64
- Haque, N., and Norgate, T., Estimation of greenhouse gas emissions from ferroalloy production using life cycle assessment with particular reference to Australia, J Cleaner Production, 2013, 39, 220-230
- Haque, N., and Norgate, T., Assessing some environmental impacts of mining, mineral processing and metal production,
 ALTA 2013 Gold Conference, 2013, 25th May-1st June,
 Perth, WA, Australia, 177-185
- Haque, N., and Northey, S., Application of life cycle assessment methodology for assessing environmental impacts of mining, mineral processing and metal production, Workshop on Life Cycle Assessment (LCA) for Mining Mineral Processing and Metal Extraction and Power Generation, 2014, 4th-5th April, Indian School of Mines, Dhanbad, India 35-44
- Hart, A., Clift, R., Riddlestone, S., and Buntin, J., Use of life cycle assessment to develop industrial ecologies - A case study graphics paper, Process Safety and Environmental Protection, 2005, 83(4), 359-363
- Hidalgo, I., Szabo, L., Ciscar, J.C., and Soria, A., *Technological* prospects and CO₂ emission trading analyses in the iron and steel industry: A global model, Energy, 2005 30(5), 583-610
- Hiraki, T., and Akiyama, T., Exergetic Life Cycle Assessment of New Waste Aluminium Treatment System with Co-production of Pressurised Hydrogen and Aluminium Hydroxide, Intl. J Hydrogen Energy, 2009, 34, 153-161
- Hiraki, T., Yamuchi, S., Iida, M., Uesugi, H., and Akiyama, T., Process for Recycling Waste Aluminium with Generation of High-Pressure Hydrogen, Environ. Sci.Technol., 2007, 41, 4454-4457
- Hischier, R., *Life cycle inventories of packagings and graphical paper*, Ecoinvent Report, 2007 Final report ecoinvent data v2.0 No. 11. Swiss Centre for Life Cycle Inventories, Dübendorf, CH.
- Ho, J.C., and Chandratilleke, T.T., Thermodynamic analysis of an electric arc furnace, Energy Conversion and Management, 1991, 31(2), 179-186

- Hu, C., Chen, L., Zhang, C., Qi, Y., Yin, R., Emission mitigation of CO₂ in steel industry: current status and future scenarios, Journal of Iron and Steel Research, International, 2006, 13(6), 38-42
- Hu, S.D., and Zandi, I., *The economics of energy conservation policies A study of US primary copper production*, Energy Economics, 1979, 173-179
- Hydro, Norway, Aluminium, environment and society, 2012, 1-56
- Institute of Scrap Recycling Industries Inc., *Comments* submitted to Proposed Rule Amending the Definition of Solid Waste, 2011, 1-175, (Docket ID No. EPA-HQ-RCRA-2010-0742)
- International Aluminium Institute Bauxite & Alumina Committee, *Alumina Technology Roadmap*, 2006, 1-52
- International Aluminium Institute, *Global Life Cycle Inventory*Data for the Primary Aluminium Industry 2010 Data, World Aluminium, 2013, 1-53
- International Aluminium Institute, Results of the 2013 Anode effect Survey Report on the Aluminium Industry's Global Perfluorocarbon Gases Emissions Reduction Programme, World Aluminium, 2014, 1-26
- International Aluminium Institute, *The Global Aluminium Industry 40 years from 1972*, World Aluminium, 2013, 1-27
- International Copper Association Copper Recycling, 2014, 1-4
- International Copper Study Group, The World Copper Factbook, 2014, 1-63
- International Energy Agency, Tracking Industrial Energy Efficiency and CO₂ Emissions, 2007, 1-234
- International Energy Agency, IEA Statistics CO₂ Emissions from Fuel Combustion Highlights, 2013, 1-158
- International Energy Agency, IEA Statistics CO₂ Emissions from Fuel Combustion Highlights, 2014, 1-136
- International Iron Metallics Association, [cited 08/08/15] Available from: http://metallics.org.uk/
- Ioana, A., and Semenescu, A., Technological, Economic, and Environmental Optimization of Aluminium Recycling, JOM, 2013, 65(8), 951-957
- Johnson, J., Reck, B.K., Wang, T and Graedel, T.E., The Energy Benefit of Stainless Steel Recycling, Energy Policy, 2008, 36, 181-192
- Jolly, J.L., The U.S. Copper-base Scrap Industry and Its Byproducts An Overview 2013 Technical Report Thirteenth Edition, Copper Development Association Inc., 2013, 1-106
- Kellogg, H.H., Sizing up the energy requirements for producing primary metals, Engineering and Minerals Journal, 1977, 178(4), 61-65
- Kim, H.C., and Wallington, T.J., Life-cycle energy and greenhouse gas emission benefits of lightweighting in automobiles: Review and harmonization, Environ. Sci. Technol., 2013, 47(12), 6089-6097
- Kim, Y., and Worrell, E., International comparison of CO₂ emission trend in the iron and steel industry, Energy Policy, 2002, 30, 827-838
- Koltun, P., Tharumarajah, A., and Grandfield, J.F., *Greenhouse* emissions in primary aluminium smelter cast houses A life cycle analysis, Materials Science Forum, 2009, 630, 27-34
- Kuckshinrichs, W., Zapp, P., and Poganietz, W-R., CO₂ emissions of global metal industries: the case of copper, Applied Energy, 2007, 84, 842-852

- Leach, M.A., Bauen, A., and Lucas, N.J.D., A Systems Approach to Materials Flow in Sustainable Cities: A Case Study of Paper, Journal of Environmental Planning and Management, 1997, 40(6), 705-724
- Life Cycle Assessment of aluminium: inventory data for the primary aluminium industry. Year 2005 update. International Aluminium Institute. September 2007
- Lipowsky, H., and Arpaci, E., *Copper in the Automotive Industry*, Wiley-VCH Verlag GmBH & Co, KgaA. Weinheim ISBN: 978-3-527-31769-1
- Liu, G., and Muller, D.B., Addressing Sustainability in the Aluminium Industry: A Critical Review of Life Cycle Assessments, J Cleaner Production, 2012, 35, 108-117
- Liu, G., and Muller, D.B., Mapping the Global Journey of Anthropogenic Aluminium: A Trade-linked Multilevel Material Flow Analysis Environ.Sci.Technol., 2013, 47(20), 11873-11881
- Logozar, K., Radonjic, G., and Bastic, M., *Incorporation of reverse logistics model into in-plant recycling process: A case of aluminium industry*, Resources, Conservation and Recycling, 2006, 49(1) 49-67
- Lucio, N.R., de Queiroz Lamas, W., and de Camargo, J.R., Strategy Energy Management in the Primary Aluminium Industry: Self-generation as a Competitive Factor, Energy Policy 2013, 59, 182-188
- Luttropp, C., and Johanson, J., Improved recycling with life cycle information tagged to the product, J Cleaner Production, 2010, 18(4), 346-354
- Magnaghi, G., *Recovered Paper Market in 2012*, Bureau of International Recycling, 2014, 1-24
- McLellan, B.C., Corder, G.D., Giurco, D.P., and Ishihara, K.N., Renewable energy in the minerals industry: a review of global potential, J Cleaner Production, 2012, 32, 32-44
- McMillan, C.A., and Keoleian, G.A., *Not All Primary Aluminium* is Created Equal: Life Cycle Greenhouse Gas Emissions from 1990 to 2005, Environ.Sci.Technol., 2009, 43, 1571-1577
- Memary, R., Giurco, D., Mudd, G., Mason, L., Life cycle assessment: a time-series analysis of copper, J.Cleaner Production, 2012, 33, 97-108
- Menzie, W.D., Barry, J.J., Beliwas, D.I., Bray, E.L., Goonan, T.G., and Matos, G, *The Global Flow of Aluminium from 2006 through 2025*, US Geological Survey, 2010, 1-78
- Midrex Technologies, Inc., 2013 World Direct Reduction Statistics Audited by World Steel Dynamics, 2014, 1-14
- Milford, R.L., Allwood, J.M., and Cullen, J.M., Assessing the Potential of Yield Improvements Through Process Scrap Reduction for Energy and CO₂ Abatement in the Steel and Aluminium Sectors Resources, Conservation and Recycling, 2011, 55, 1185-1195
- Monte, M.C., Fuente, E., Blanco, A., and Negro, C., *Waste management from pulp and paper production in the European Union*, Waste Management, 2009, 29(1), 293-308
- Morris, J., Recycling vs. incineration: an energy conservation analysis, J. Hazardous Materials, 1996 47, 277-293
- Muir, D.M., The Parker Copper Process a new approach ahead of its time, Hydrometallurgy 2008: Proceedings of the 6th International Symposium, 17th-20th August, Phoenix, Arizona, USA

- Moskalyk, R.R., and Alfantazi, A.M., *Review of copper pyrometallurgical practice: today and tomorrow*, Minerals Engineering 2003, 16, 893–919
- Najdenov, I., Raić, K.T., and Kokeza, G., Aspects of energy reduction by autogenous copper production in the copper smelting plant Bor, Energy, 2012, 43, 376-384
- Nakajima, K., Osuga, H., Yokoyma, K., and Nagasaka, T., Material flow analysis of aluminium dross and environmental assessment for its recycling process, Journal of the Japan Institute of Metals, 2008, 72(1), 1-7
- Norgate, T. E., Metal recycling: An assessment using life cycle energy consumption as a sustainability indicator, CSIRO Minerals (Report), 2004, 1-44
- Norgate, T.E., Assessing the sustainability of aluminium and steel production using exergetic life cycle assessment, 6th Annual Conference on Life Cycle Assessment Sustainability Tools for a New Climate, 2009, 16th-19th February, Melbourne, Victoria, Australia
- Norgate, T and Haque, N., Energy and Greenhouse Gas Impacts of Mining and Mineral Processing Operations, J.Cleaner Production, 2010, 18, 266-274
- Norgate, T. E., and Rankin, W. J., *The role of metals in sustainable development*, Green Processing, (The AusIMM), Caims, 2002, 49-55
- Norgate, T., and Jahanshahi, S., Reducing the Greenhouse Gas Footprint of Primary Metal Production: Where should the focus be? Minerals Engineering, 2011, 24, 1563-1570
- Norgate, T., and Langberg, D., *Environmental and economic aspects of charcoal use in steelmaking*, ISIJ International, 2009, 49(4), 587-595
- Norgate, T.E., and Rankin, W.J., *Greenhouse Gas Emissions from Aluminium Production A Life Cycle Approach*CSIRO Minerals, www.minerals.csiro.au/sd/CSIRO_Paper_LCA_Al.htm
- Norgate, T.E., and Rankin, W.J., *Life cycle assessment of copper and nickel production*, Proceedings, Minprex 2000, International Conference on Minerals Processing and Extractive Metallurgy, 2000, 133-138
- Norgate, T.E., Jahanshahi, S., and Rankin, W.J., Assessing the Environmental Impact of Metal Production Processes J Cleaner Production, 2007, 15, 838-848
- Norgate, T.E., Jahanshahi, S., and Rankin, W.J., *Alternative* routes to stainless steel A life cycle approach, CSIRO Minerals Proceedings: Tenth International Ferroalloys Congress, 2004, 1st–4th February, Cape Town, South Africa, 693-704
- Northey, S., and Haque, N., Assessment of greenhouse gas emissions from copper production: A case study of El Teniente Copper Mine, Environmene 2013 3rd International Seminar on Environmental Issues in the Mining Industry, 2013, 4th-6th December, Santiago, Chile, 99-109
- Northey, S., Mohr, S., Mudd, G., Weng, Z., and Giurco, D., Modelling future copper ore grade decline based on a detailed assessment of copper resources and mining, Resources, Conservation and Recycling, 2014, 83, 190-201
- Northey, S., Haque, N., and Mudd, G., *Using sustainability reporting to assess the environmental footprint of copper mining*, J. Cleaner Production, 2013, 40, 118-128

- Olivieri , G., Romani, A., and Neri, P., *Environmental and Economic Analysis of Aluminium Recycling Through Life Cycle Assessment*, Intl.J.Sustainable Development & World Ecology, 2006, 13, 269-276
- Onyedika, G.O., Achusim-Udenko, A.C., Nwoko, C.I.A., Ogwuegbu, M.O.C., Chemistry, processes and problems of complex ores utilisation: hydrometallurgical options, Int. J. Chem. Sci., 2012, 10(1), 112-130
- Orth, A., Anastasijevic, N., and Eichberger, H., Low CO₂ emission technologies for iron and steelmaking as well as titania slag production, Minerals Engineering, 2007, 20, 854-861
- Ozawa, L., Sheinbaum, C., Martin, N., Worrell, E., and Price, L. Energy use and CO₂ emissions in Mexico's iron and steel industry, Energy, 2002, 27, 225-239
- Pulp & Paper Resources & Information Site, *Country wise paper and paperboard production & consumption statistics*, [cited: 18/03/15], Available from: http://www.paperonweb.com/
- Paper Recycling Association of South Africa, *Paper recycling statistics for 2013*, [cited: 12/02/15] Available from: http://www.prasa.co.za/statistics
- Paraskevas, D., Kellens, K., Renaldi, Dewulf, W., and Duflou, J.R., Closed and Open Loop Recycling of Aluminium: A Life Cycle Assessment Perspective, Proceedings of the 11th Global Conference on Sustainable Manufacturing Innovative Solutions, 2013, 23rd-25th September Berlin, Germany, 305-310
- Passarini, F., Ciacci, L., Santini, A., Vassura, I., and Morselli, L., *Aluminium flows in vehicles: enhancing the recovery at end-of-life*, J Mater Cycles Waste Manag, 2014, 16, 39-45
- Pati, R.K., Vrat, P., and Kumar, P., A goal programming model for paper recycling system, Omega, 2008, 36(3), 405-417
- Phylipsen, D., Blok, K., Worrell, E., and Beer, J., *Benchmarking* the energy efficiency of Dutch industry: an assessment of the expected effect on energy consumption and CO₂ emissions, Energy Policy, 2002, 30, 663-679
- Polinares Working Paper No2, Copper, 2012
- Polinares Working Paper No40, Factsheet: Copper, 2012, 1-18
- Price, L., Sinton, J., Worrell, E., Phylipsen, D., Xiulian, H., and Ji, L. *Energy use and carbon dioxide emissions from steel production in China*, Energy, 2002, 27 429-446
- Quinkertz, R., Rombach, G., and Liebig, D., A scenario to optimise the energy demand of aluminium production depending on the recycling quota, Resources, Conservation and Recycling, 2001, 33, 217-234
- Rankin, J., *Energy Use in Metal Production*, High Temperature Processing Symposium, Swinburne University of Technology, 2012, 7-9
- Rubach, C., World Steel Recycling in Figures 2009 2013 Steel Scrap – a Raw Material for Steelmaking, Bureau of International Recycling, 2014, 1-36
- Ruth, M., Dematerialization in five US metals sectors: implications for energy use and CO₂ emissions, Resources Policy., 1998, 24(1), 1-18
- Rynikiewicz, C., The climate change challenge and transitions for radical changes in the European steel industry, J Cleaner Production, 2008, 16(7), 781-789

- Sakamoto, Y., Tonooka, Y., and Yanagisawa, Y., Estimation of Energy Consumption for Each Process in the Japanese Steel Industry: A Process Analysis, Energy Conversion & Management, 1999, 40, 1129-1140
- Sakamoto, Y. and Tonooka, Y., Estimation of CO₂ emission for each process in the Japanese steel industry: a process analysis, International Journal of Energy Research, 2000, 24, 625-632
- Samuel, M., A New Technique for Recycling Aluminium Scrap J Mater. Processing Technology 2003, 135, 117-124
- Schenk, N.J., Henri, C.M., and Potting, J., *The nonlinear relationship between paper recycling and primary pulp requirements*, Journal of Industrial Ecology, 2004, 8(3), 141-161
- Schlesinger, M.E., Aluminium Recycling 2nd Edition, CRC Press Taylor and Francis Group, 2014, ISBN: 13:978-1-4665-7025-2
- Schmidt, J.H., Holm, P., Merrild, A., and Christensen, P., *Life* cycle assessment of the waste hierarchy a Danish case study on waste paper, Waste Management, 2007, 27, 1519-1530
- Schwarz, H-G., Briem, S and Zapp, P., Future Carbon Dioxide Emissions in the Global Material Flow of Primary Aluminium, Energy, 2001, 26, 775-795
- Schwarz, H-G., Technology Diffusion in Metal Industries: Driving Forces and Barriers in the German Aluminium Smelting Sector, J Cleaner Production, 2008, 16S1, S37-S49
- Shao, C., Guan, Y., Wan, Z., Chu, C., and Ju, M., Performance Analysis of CO₂ Emissions and Energy Efficiency of Metal Industries in China, J.Environ.Management, 2014, 134, 30-38
- Song, X., Yang, J., Lu, B., and Li, B., Exploring the life cycle management of industrial solid waste in the case of copper slag, Waste Management & Research, 2013, 31(6), 625-633
- Song, X., Yang, J., Lu, B., Li, B., and Zeng, G., Identification and Assessment of Environmental Burdens of Chinese Copper Production from a Life Cycle Perspective, Front.Environ.Sci. Eng., 2014, 8(4), 580-588
- Spatari, S., Bertram, M., Fuse, K., Graedel, T.E., Rechberger, H., The contemporary European copper cycle: 1 year stocks and flows, Ecological Economics, 2002, 42, 27-42
- Statistics and facts about the global paper industry, [cited: 18/03/15] Released 2013: Available from: http://www.statista.com/topics/1701/paper-industry
- Steel Times International A Report on the Environmental Benefits of Recycling A Critical Review of the Data for Steel 2010
- Swart, P., and Dewulf, J., Modeling Fossil Energy Demands of Primary Nonferrous Metal Production: The Case of Copper, Environ. Sci. Technol., 2013, 47, 13917–13924
- The Aluminium Association, Aluminium: The Element of Sustainability, 2011, 1-70
- The Aluminium Association, The Environmental Footprint of Semi-Finished Aluminium Products in North America – A Life Cycle Assessment Report, 2013, 1-124
- The Metrics of Material and Metal Ecology Harmonizing the Resource, Technology Chapter 13 Aluminium Metal Production, Developments in Mineral Processing, 2005, 16, 391-451, Elsevier 2015, ISBN: 978-0-444-51137-9
- Udo de Haes, H.A., and Heijungs, R., *Life cycle assessment for energy analysis and management*, Applied Energy, 2007, 84(7), 817-827

- US Energy Information Administration, Today in Energy Energy needed to produce aluminium, 2012, http://www.eia.gov/todayinenergy/
- US Energy Information Administration, Today in Energy Recycling is the primary energy efficiency technology for aluminium and steel manufacturing, 2014, http://www.eia.gov/todayinenergy/
- US Energy Information Administration, Manufacturing Energy Consumption Survey, 2013 http://www.eia.gov/consumption/manufacturing/index.cfm
- US Geological Survey, *Mineral Commodity Summaries*, 2014, 1-199, Available from: http://minerals.usgs.gov/minerals/pubs/commodity/
- Villanueva, A., and Wenzel, H., Paper waste Recycling, incineration or landfilling? A review of existing life cycle assessments, Waste Management, 2007, 27(8), S29-S46
- Villar, A., Arribas, J.J., and Parrondo, J., Waste-to-energy technologies in continuous process industries, Clean Techn Environ Policy, 2012, 14, 29-39
- Wang, K., Wang, C., Lu, X., and Chen, J. Scenario analysis on CO₂ emissions reduction potential in China's iron and steel industry, Energy Policy, 2007, 35, 2320-2335
- Welch, B.J., Hyland, M.M., and James, B.J., Future Materials Requirements for the High-Energy-Intensity Production of Aluminium JOM, 2001, 53(2), 13-18
- Wei, W., Energy consumption and carbon footprint of secondary aluminium cast house, Master Thesis, Royal Institute of Technology, Sweden
- Wei, Y.M., Liao, H., and Fan, Y., An empirical analysis of energy efficiency in China's iron and steel sector, Energy, 2007, 32, 2262-2270
- World Steel Association, Fact Sheet Energy Use in the Steel Industry, 2014, 1-3
- World Steel Association, Fact Sheet Steel and Raw Materials, 2014, 1-2
- World Steel Association, *World Steel in Figures 2014*, 2014, 1-17 [cited 12/02/15] Available from: http://www.worldsteel.org/statistics/economics.html
- Worrell, E., Price, L., Martin, N., Farla, J., and Schaeffer, R., Energy intensity in the iron and steel industry: a comparison of physical and economic indicators, Energy Policy, 1997, 25(7-9), 727-744
- Worrell, E., Price, L., Neelis, M., Galitsky, C., and Zhou, N., World Best Practice Energy Intensity Values for Selected Industrial Sectors Published in the United States by Environmental Energy Technologies Division Lawrence Berkeley National Laboratory, Berkeley, USA, 2007, 1-50
- Xiao, Y., and Reuter, M.A., *Recycling of Distributed Aluminium Turning Scrap*, Minerals Eng. 2002, 15, 963-970
- Xueyi, G., and Yu, S., *Substance flow analysis of copper in China*, Resources, Conservation and Recycling, 2008, 52(6), 874-882
- Yanjia, W., and Chandler, W., The Chinese Nonferrous Metals Industry – Energy Use and CO₂ Emissions Energy Policy, 2010, 38, 6475-6484
- Zhang, W., Li, H., Chen, B., Li, Q., Hou, X., and Zhang, H., CO₂ Emission and Mitigation Potential Estimations of China's Primary Aluminium Industry J Cleaner Production, 2014, 1-10 (doi:10.1016/j.jclepro.2014.07.066)

Annex

Questions Prepared for BIR to put to Members

- 1. What metals/materials do you recover in the plant?
- 2. What is the tonnage of feedstock through the plant per day?
- 3. What is the typical percentage of aluminium, copper and ferrous metals in the feedstock? Or paper?
- 4. What is the total energy consumption through the plant per day? [If possible provide an estimate of the energy sources e.g. oil, gas, electricity, etc]
- 5. What is the tonnage of recovered metal at the end of the process?

- 6. Does the recovered metal require further refinement after it leaves the plant?
- 7. If you recover more than one metal, provide an estimate of the fraction of energy usage that you believe is attributable to the recovery of aluminium, copper or ferrous metal fractions.
- 8. It would be helpful if you could indicate which of the following regions you operate in: (a) Europe; (b) North America; (c) South America; (d) Africa; (e) Asia; (f) Australasia.





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